

1 Tests of an experimental unbaited video lander as a marine fish survey tool for high-relief
2 deepwater rocky reefs

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4 Robert W. Hannah*

5 Matthew T. O. Blume

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12 Oregon Department of Fish and Wildlife

13 Marine Resources Program

14 2040 SE Marine Science Drive

15 Newport, Oregon 97365, U.S.A.

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17 *Corresponding author: Tel.: 541 867-0300 ext. 231, Fax: 541 867-0311, email:

18 bob.w.hannah@state.or.us

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20 Key words: sampling methods, untrawlable habitat, underwater visual census, marine

21 protected areas, yelloweye rockfish, temperate reefs

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23

24 **Abstract**

25 We describe the development and practical field testing of a rugged, unbaited video
26 lander as a visual survey tool for fishes inhabiting structurally complex, high-relief,
27 deepwater rocky reefs. Our autonomous, high-resolution, low-light, color video lander
28 system utilized a smooth frame design that incorporated a series of breakaway
29 attachments and inexpensive sacrificial steel bases to maximize the potential for camera
30 system recovery from complex rocky habitats. Initial field tests at five reef complexes
31 off Oregon (n=421) and a larger study evaluating the western boundary of Oregon's
32 Yelloweye Rockfish Conservation Area (YRCA, n=527) showed that the video lander
33 could be deployed and reliably retrieved from high-relief rocky habitat without damage to
34 the camera system and with minimal losses of sacrificial bases. Acceptable visibility for
35 counting fish from the lander video was common at offshore reefs like Stonewall Bank,
36 but less so at nearshore reef complexes. The video lander system was effective for
37 discriminating differences in fish species assemblages at the various reefs surveyed (one-
38 way ANOSIM, $P < 0.001$) and for identifying seafloor habitat types and species-habitat
39 associations for yelloweye rockfish (*Sebastes ruberrimus*) at Stonewall Bank ($P < 0.05$).
40 The video lander data showed that the area outside and to the west of the YRCA enclosed
41 similar quantities of yelloweye rockfish and their preferred habitats in comparison with
42 the area inside the YRCA (0.179 yelloweye rockfish/station outside, versus 0.144
43 yelloweye rockfish/station inside, Wilcoxon test, $P = 0.417$). Our visual survey data also
44 showed that the current western YRCA boundary is not optimal for protecting yelloweye
45 rockfish at Stonewall Bank from fishery harvest.

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47

48 **1.0 Introduction**

49 A variety of tools have been developed for conducting visual surveys of marine fish
50 populations inhabiting rocky reefs. The various tools have different biases (e.g. St. John
51 et al. 1990, Willis et al. 2000, Harvey et al. 2007, Stoner et al. 2008) but all offer the
52 benefit of non-extractive sampling, which can be important for sampling in marine
53 reserves, marine protected areas (MPAs) and untrawlable habitats. Baited and unbaited
54 underwater video stations and direct visual and video surveys by scuba divers have all
55 been used successfully to survey fish populations on shallow-water (approximately ≤ 20
56 m) reefs (e.g. Willis and Babcock 2000, Watson et al. 2005, Langlois et al. 2010). Baited
57 and unbaited video stations or “landers” have also been used with success in much deeper
58 waters (e.g. Ellis and DeMartini 1995, Harvey et al. 2007).

59

60 Visual fish surveys of structurally complex, high-relief deepwater reefs have also been
61 conducted, usually with more expensive and technically sophisticated equipment such as
62 remotely operated vehicles (ROVs) or human occupied vehicles (HOVs, O’Connel and
63 Carlile 1993, Yoklavich et al. 2007, Stoner et al. 2008). In addition to being expensive to
64 purchase or lease, these vehicles require highly skilled operators. The ships that are
65 needed to safely deploy all but the smallest of these types of survey vehicles can also be
66 prohibitively expensive for small-scale surveys. Recently, very heavily constructed
67 towed stereo-camera sleds have been used with some success to survey these deep,
68 rugose habitats with less costly equipment (Williams et al. 2010). We report here on the
69 development and extensive field testing of another type of relatively inexpensive visual
70 survey tool for these structurally-complex habitats, specifically an unbaited video lander

71 that was purpose-built for deployment into very rugged habitats. The objectives of our
72 study were to:

73

74 1) Develop a design for a video lander that could be deployed and reliably retrieved,
75 without damage, from high-relief, deepwater rocky reefs.

76

77 2) Determine, through field testing at a variety of reefs and across different conditions, if
78 a video lander could be effectively used to gather data on differences in reef fish species
79 assemblages in U.S. Pacific northwest marine waters.

80

81 3) Evaluate the video lander further by using it in an applied, large-scale field study to
82 measure spatial differences in the relative abundance of fishes.

83

84 **2.0 Methods**

85 *2.1 Lander design*

86 Our autonomous, high-resolution, low-light, color video lander system was purpose-built
87 for grid-based visual surveys of very high-relief rocky habitat (Figure 1). It was designed
88 to be deployed and retrieved from vessels equipped with just a hydraulic crab block, and
89 to be very portable, to facilitate use during brief periods of good ocean conditions. The
90 concept we worked from was to have a relatively inexpensive system with no video feed
91 to the surface, designed to be deployed to the bottom for a fixed time interval at a series
92 of sites in a sequence, much like launching and retrieving a crab pot. We chose not to use
93 bait with our video system so that fish presence/absence data could be used to analyze
94 species-habitat associations. Baited underwater video systems (BUVs) have been shown

95 to improve statistical power of survey data for some species due to the fact that they
96 attract more fish, presumably from a greater distance (Harvey et al. 2007). However, this
97 effect is also likely to bias species-habitat associations.

98

99 Our design for the frame and buoy line incorporated several features to increase the
100 likelihood of recovery of the system in the event the lander became stuck in rocky
101 habitat. The frame was constructed to be very smooth in shape to avoid hangups on rock
102 ledges and strong enough to survive contact with rock surfaces without damage (Figure
103 1). The frame was constructed in a cage-like shape to provide protection for the camera,
104 lights, and pressure housing containing the batteries and recording device (Figure 1). It
105 incorporated an inexpensive sacrificial base constructed of 13 mm diameter mild steel
106 rod, designed to break away if necessary, freeing the upper portion of the lander and the
107 video equipment to return to the surface. The sacrificial base was weighted with one or
108 two (more weight was used in deeper water) 9.1 kg, 61 mm lengths of 51 mm diameter
109 mild steel rod to help maintain a vertical orientation during descent to the seafloor. The
110 bottom of the base was covered with lightweight netting material to keep pointed rocks
111 from protruding up into the lander frame and contacting the video equipment. The buoy
112 line to the surface was constructed mostly of floating line to minimize fouling in rocks.
113 A 4.5 m length of abrasion-resistant 9.5 mm line was used at the terminal end of the buoy
114 line to limit abrasion by rocks and debris. On the lander frame a “breakaway” line
115 constructed of 8 mm Spectra[®] was attached in a way such that as increased pulling force
116 was applied to the line, several “weak links” would break sequentially, causing the lander
117 to tip, and then rotate, to hopefully free it from any obstructions that it was hung up on
118 (Figure 1). Fully assembled, as shown in Figure 1, our video lander weighed 42.6 kg.

119

120 *2.2 Video camera system*

121 We chose a Deep Sea Power and Light (DSPL) color video camera, paired (Multi-
122 SeaCam[®] 2060) with two DSPL LED Mini-SeaLites[®] (850 lm, 6500° K) for illumination
123 (Table 1). A color camera was chosen to aid in species identification. Video was
124 captured using a SONY[®] DCR-TRV11 digital camcorder. The system utilized a pressure
125 housing constructed from 10.2 cm ID aluminum tube to enclose the batteries and
126 camcorder. Systems utilizing high-quality, consumer-level high-definition camcorders
127 can also be built for a similar total cost (Table 1). Our system was powered by two 13.2
128 volt rechargeable NiMH battery packs, wired in parallel (Table 1). The system was
129 activated at depth by a pressure-activated switch located on the end of the pressure
130 housing. When the switch was activated, a programmable micro-controller circuit board
131 conveyed power to the camera, lights and camcorder, and issued LANC (Local
132 Application Control Bus System) control commands to the camcorder to initiate video
133 recording.

134

135 Deployment and retrieval of the video lander was straightforward. Upon reaching each
136 station, the lander was simply tossed over the side of the vessel and allowed to sink to the
137 bottom. The float for the buoy line was also immediately thrown overboard and the
138 vessel drifted away from the site. After enough time had elapsed for a pre-determined
139 duration of video recording on the seafloor, the floating buoy was retrieved and the
140 lander was hauled to the surface with the vessel's hydraulic crab block.

141

142 *2.3 Initial field tests*

143 We field tested the video lander to determine if it was reliably retrievable, to evaluate the
144 frequency with which acceptable visibility and unobstructed views were obtained and to
145 see if fish noticeably avoided or were attracted to it. To accomplish this, we deployed it
146 at a variety of different reef complexes (Figure 2) encompassing a range of depths and
147 across many different months of the year. A pre-designed set of stations (mostly
148 following a rectangular grid) was utilized for most surveys although some experimental
149 deployments or “drops” (hereafter drops) at small sites known to harbor demersal fish
150 were also conducted. All surveys were conducted during daylight hours, avoiding
151 crepuscular periods during which many demersal fish are known to move off bottom
152 (Engås and Soldal 1992, Hannah et al. 2005). At Cape Perpetua, where the amount of
153 rocky reef habitat is small and surrounded by expanses of sand and gravel, we conformed
154 the station layout to fit the contours of the reefs. For all the pre-designed surveys, we
155 chose a minimum distance between stations of at least 100 m to be sure that the same fish
156 was not counted at adjacent stations. After initial testing with a range of bottom times,
157 we chose a standardized bottom time of 4-5 minutes. This bottom interval was chosen
158 based on several factors. First, a review of the initial video deployments showed that this
159 was usually enough time for any disturbed sediment to clear. Also, review of the video
160 footage from preliminary drops showed that after about 4 minutes, enough fish had
161 frequently been seen such that additional specimens could not be confidently counted as
162 fish that had not already been observed. This shorter interval also allowed for more drops
163 to be completed in a day, with a larger total area surveyed.

164

165 *2.4 Video analysis*

166 The video footage from each deployment or “drop” was first imported into Adobe
167 Premiere Pro[®] so that it could be easily reviewed on a full size computer screen. Each
168 drop was scored for visibility and the quality of the view of the surrounding habitat,
169 following the criteria shown in Table 2. Fish that were observed at each station, and that
170 could be positively identified to species or a species group, were counted based on a
171 maximum count of that species or group in any single video frame, following Harvey et
172 al. (2007). This approach resulted in conservative counts, but is necessary because it was
173 not possible to tell if a fish newly entering the frame was a different individual than one
174 seen seconds or minutes before. The primary habitat (most common habitat in the view)
175 and secondary habitat (second most common habitat type in the view) types were
176 classified for each drop, based on the criteria shown in Table 3.

177

178 *2.5 Species assemblages*

179 To evaluate the effectiveness of the video lander system for studying differences in fish
180 species assemblages, we compared the species composition at different reefs using a one-
181 way ANOSIM, as implemented in PAST 2.14 (Hammer et al. 2001). Prior to statistical
182 analysis, the count data were standardized to percentages to reduce effects from varying
183 numbers of stations at different reefs (Sommerfield 2008) and the analysis was restricted to
184 just the 16 most abundant species across all surveys. With the exception of unidentified
185 juvenile rockfish (*Sebastes* spp.), unidentified fish were excluded from the analysis. We
186 used pairwise ANOSIMS as post-hoc tests and adjusted these for multiple comparisons
187 using a step-down sequential Bonferroni correction. The degree of difference in species
188 composition between specific reef surveys was measured using the Bray-Curtis

189 dissimilarity index (BCDI, Bray and Curtis 1957), as implemented in the SIMPER
190 routine in PAST 2.14.

191

192 2.6 Large-scale field test

193 To further test the utility of our video lander for quantitative fish surveys, we employed it
194 in a large-scale field experiment to compare the relative abundance of selected fish
195 species in two areas. As our question of interest, we chose to evaluate the boundaries of
196 an MPA that was established to help rebuild depleted populations of yelloweye rockfish
197 (*Sebastes ruberrimus*, PFMC 2011). Specifically, we chose to determine if the current
198 western boundary of the northern portion of Oregon's Yelloweye Rockfish Conservation
199 Area at Stonewall Bank (YRCA, Figure 3) had been chosen well to minimize fishery
200 impacts on yelloweye rockfish. The YRCA is closed to recreational fishing for Pacific
201 halibut (*Hippoglossus stenolepis*), much of which is centered around Stonewall Bank,
202 where yelloweye rockfish can be a common bycatch species. At the time the boundaries
203 of the YRCA were first developed, very little information on the spatial distribution of
204 yelloweye rockfish or their preferred habitats at Stonewall Bank was available. Due to
205 this lack of information, the boundaries were chosen to simply enclose a large portion of
206 the shallowest parts of Stonewall Bank. The long-term effectiveness of the YRCA will
207 be determined by how completely it encloses the locations at Stonewall Bank that are
208 routinely inhabited by yelloweye rockfish. We chose the northwestern portion of the
209 YRCA for further study because of reports by some anglers that they encountered
210 yelloweye rockfish in the area immediately outside and west of the YRCA. Our goal was
211 to measure yelloweye rockfish relative abundance on both sides of the current western

212 YRCA boundary to evaluate whether the current boundary was optimal, or if a different
213 boundary, farther west, would better protect yelloweye rockfish.

214

215 We chose a sampling design with a grid spacing of 200 m to allow very intensive
216 sampling of the study sites (Figure 3) but still maintain the independence of adjacent fish
217 counts, and again used a bottom time of 4-5 minutes. To control for differences between
218 sampling days in visibility or other environmental conditions, each day of sampling was
219 divided roughly in half between stations within and outside of the YRCA (Figure 3).

220 Video footage was reviewed in the field whenever possible and stations with an
221 obstructed view of surrounding habitat were repeated as soon as possible.

222

223 Since an optimal revision of the YRCA western boundary requires a knowledge of both
224 the distribution of yelloweye rockfish and their preferred habitats, we evaluated the
225 distribution of habitat types and the relative abundance of yelloweye rockfish between the
226 two areas as well as the association of yelloweye rockfish with the primary and secondary
227 habitat types observed. For the latter analysis we used Fisher's exact test, applied to
228 presence/absence data for yelloweye rockfish. We compared the relative abundance
229 (maximum count per station) of yelloweye rockfish between the two study sites using the
230 nonparametric Wilcoxon test. Although our primary focus was on yelloweye rockfish,
231 we also present relative abundance data and Wilcoxon tests for four other common
232 demersal fish species from these two study sites to illustrate the approximate differences
233 in relative abundance that we were able to differentiate with video lander data at this reef.

234

235 **3.0 Results**

236 3.1 *Initial Field Tests*

237 Between October 2009 and September 2010, we deployed the video lander in pre-
238 designed or exploratory surveys at five different reef complexes (Figure 2), usually
239 sampling between 30 and 40 stations per day. In the course of 421 initial drops, the video
240 system was never lost or damaged. During four drops, the sacrificial base broke free and
241 was lost after the lander became badly stuck in rocky habitat, showing the benefit of this
242 element of our design. The breakaway attachments on the buoy line were also frequently
243 broken when the lander got temporarily stuck, suggesting that this element of the design
244 also contributed to successful retrieval of the video system from rocky reef habitats.
245 Strong currents at Stonewall Bank also sank the surface buoy on one occasion; however,
246 we were able to successfully grapple for the buoy line and retrieve it. The addition of a 3
247 m section of line with a trailing buoy attached to the primary surface buoy resolved this
248 problem.

249

250 3.2 *Effectiveness for visual surveys*

251 The video lander system was mostly very effective at obtaining video footage of the
252 survey stations at the five reef complexes initially sampled. On a few drops, the
253 pressure-activated switch failed to initiate video recording, a problem that was fixed with
254 some simple adjustments. Excluding drops that produced no video footage (n=16),
255 moderate or good visibility (Table 2) for counting fish was encountered in 89% of drops.
256 At offshore stations like Stonewall Bank, good visibility was typical, with only 1.7% of
257 drops rated as poor. On some days at Stonewall Bank, visibility was so good that
258 significant amounts of surface light reached the seafloor, causing increased detection
259 distance for fish, but also shifting the video to a bluish shade. At nearshore stations,

260 visibility was more frequently rated as poor and was much more variable, especially at
261 sites closer to shore. For the four nearshore reef areas, 18.2% and 53.3% of video drops
262 were scored as having poor or moderate visibility, respectively. A completely obstructed
263 or partially obstructed video view (Figure 4) was obtained in 4.8% and 5.7%,
264 respectively, of the lander drops. Obstructed views were easily remedied by simply
265 resampling those stations while still in the field, however, poor visibility sometimes
266 required the video lander surveys to be halted and rescheduled for a later date.

267

268 3.3 Species Assemblages

269 Fish counts from the initial pre-designed video lander surveys (Table 4) showed distinct
270 differences in species assemblages between reef areas ($P < 0.001$, one-way ANOSIM,
271 Table 4). One survey site, located in the northwest portion of Siletz Reef (not shown),
272 was determined from the video lander footage to be mostly sand habitat intermixed with a
273 few small patches of rocky reef, and was excluded from further analysis. Interestingly,
274 this area was chosen for a survey because it had been identified as probable rock habitat
275 based on interpretation of bathymetric and other available data by Romsos et al. (2007).
276 The most distinct difference in species assemblages between reefs was found for
277 Stonewall Bank in comparison with the Eastern Siletz site (Figure 2, $BCDI = 91.9$,
278 $P < 0.05$, Table 4), areas with very different mean depths. The difference in assemblages
279 was driven mostly by differences in the relative abundance of canary rockfish (*S.*
280 *pinniger*) and kelp greenling (*Hexagrammos decagrammus*). The species assemblage at
281 Seal Rocks reef was distinct from all other sites ($BCDI > 85.1$, $P < 0.05$), due primarily to
282 the strong dominance of black rockfish (*S. melanops*) and relatively low abundance of
283 blue (*S. mystinus*) and canary rockfish observed at this site (Table 4). The repeat surveys

284 at Cape Perpetua reef produced roughly similar species assemblages (BCDI=78.1,
285 $P>0.05$), that differed sharply from the assemblage at Eastern Siletz reef (BCDI>89.7,
286 $P<0.05$) primarily due to a higher percentage of canary rockfish and a lower percentage
287 of blue rockfish at Cape Perpetua (Table 4). The species assemblage observed at the
288 Western Siletz site was also distinct from that found at Stonewall Bank (BCDI>77.3,
289 $P<0.05$) and Eastern Siletz (BCDI=87.1, $P<0.05$) due primarily to differences in the
290 relative abundance of canary rockfish, but also to differences in abundance of blue
291 rockfish and kelp greenling (Table 4).

292

293 3.4 *Fish behavior*

294 The large number and diversity of fish observed at the surveyed reefs (Table 4) showed
295 that most of the rocky-reef species of interest in this study did not strongly avoid the
296 video lander. Notably however, cryptic species like cabezon (*Scorpaenichthys*
297 *marmoratus*) were rarely observed (Table 4). Our observations at some stations also
298 indicated that some fish species may be attracted to the video lander. At several stations,
299 yelloweye, quillback (*S. maliger*) and copper rockfish (*S. caurinus*) were sometimes
300 observed approaching the lander from a distance.

301

302 3.5 *Large-scale field test*

303 We sampled 527 stations with the video lander at Stonewall Bank over 14 days of
304 sampling between April 19 and August 15, 2011, losing only two of the sacrificial bases
305 and with no damage to the lander or video system (Figure 3). These video lander surveys
306 also generally encountered good water clarity and unobstructed views. A total of 34
307 stations had to be sampled twice to obtain acceptable video footage for all stations.

308

309 More yelloweye rockfish were observed in the area immediately west of the YRCA than
310 in the YRCA (47 versus 38, Table 5, Figure 3). The mean number of yelloweye rockfish
311 observed was 0.179 yelloweye rockfish per station in the area west of the YRCA and
312 0.144 yelloweye rockfish per station inside the YRCA, a difference of about 24%. This
313 difference, however, was not statistically significant ($P = 0.417$, Wilcoxon test),
314 suggesting that the area outside of the YRCA to the west has an abundance of yelloweye
315 rockfish that is similar to the area enclosed within the YRCA. Lingcod (*Ophiodon*
316 *elongatus*) were about 50% less abundant ($P < 0.01$) in the area to the west of the YRCA
317 than inside it. The other differences we noted in demersal fish relative abundance were
318 not statistically significant with our sampling design (Table 5).

319

320 The frequency of primary habitat types was similar between the two sites at Stonewall
321 Bank, but did show a few key differences (Figure 5). Flat bedrock and crevice habitats
322 were more common inside the YRCA, while large and small boulder, gravel/pebble and
323 cobble habitats were more frequently observed outside the YRCA (Figure 5). Secondary
324 habitats showed a similar pattern, except that small boulder and bedrock outcrop were
325 less frequently encountered outside of the YRCA (Figure 5).

326

327 Primary and secondary habitat types were useful predictors of the presence/absence of
328 yelloweye rockfish as measured with the video lander at Stonewall Bank. Bedrock
329 outcrop as a primary habitat type was negatively associated with yelloweye rockfish
330 ($P > 0.01$). Conversely, crevice as a primary habitat type was strongly positively
331 associated with the presence of yelloweye rockfish ($P < 0.001$). Similarly, bedrock

332 outcrop as a secondary habitat type was negatively associated with yelloweye rockfish
333 ($P<0.01$), as was cobble ($P<0.05$). Large boulder as a secondary habitat type was
334 positively associated with yelloweye rockfish presence ($P<0.001$).

335

336 **4.0 Discussion**

337 The general utility of fixed underwater video stations for describing fish species
338 assemblages and measuring differences in relative abundance is well established (Willis
339 and Babcock 2000, Watson et al. 2005, Harvey et al. 2007, Langlois et al. 2010). The
340 results from our study suggest that, with a proper frame design, those capabilities can also
341 be utilized for some deepwater, highly rugose rocky reef habitats, without unacceptable
342 risk of loss or damage to the camera system. Studies comparing different visual census
343 methods for reef fishes generally recognize some bias in all techniques and suggest that
344 the choice of technique for surveys should consider the known behaviors of the species of
345 interest and the biases inherent in the technique (Willis and Babcock 2000, Watson et al.
346 2005). Our video lander system worked well for our primary species of interest at
347 Stonewall Bank, yelloweye rockfish, for several reasons. Yelloweye rockfish are a large
348 rockfish, have distinct coloration and body shape, are strongly demersally oriented and
349 are believed to be fairly sedentary, making them readily identifiable with a benthic video
350 lander employing a color camera (Love et al. 2002, Hannah and Rankin 2011). Although
351 small fish are generally harder to accurately identify from video footage, juvenile
352 yelloweye rockfish have very distinctive striped markings (Figure 4) that greatly alleviate
353 this problem (Love et al. 2002). Our observations also suggest that yelloweye rockfish
354 and most other demersal fishes inhabiting the areas we sampled did not appear to avoid
355 the video lander and that some have at least a moderate tendency to approach it,

356 potentially improving sampling efficiency for some species. However, although mild
357 attraction may increase sampling efficiency for some species, it may also bias
358 information on species composition.

359

360 Our data on yelloweye rockfish relative abundance, distribution and habitat associations
361 at Stonewall Bank provided convincing evidence that a relocation of the western
362 boundary of the YRCA further west would provide improved protection for yelloweye
363 rockfish from fishery impacts (Figure 5). The positive association we demonstrated
364 between the presence of yelloweye rockfish and crevice and large boulder habitat is
365 consistent with the findings of O'Connell and Carlile (1993) for yelloweye rockfish in the
366 eastern Gulf of Alaska. Similarly, the negative association found for this species with
367 bedrock outcrops at Stonewall Bank is consistent with the very low densities of
368 yelloweye rockfish they report for continuous rock bottom habitat (O'Connell and Carlile
369 1993). Given that yelloweye rockfish are currently considered overfished and at
370 relatively low population levels in waters off the U.S west coast (PFMC 2011), the choice
371 of an optimal new western boundary for the YRCA should consider the spatial
372 distribution of both observed yelloweye rockfish and their preferred habitats at Stonewall
373 Bank (Figure 6).

374

375 The video lander system developed here adds to the tools available for studying
376 deepwater rocky reef fishes, however, due to the very limited amount of habitat viewed in
377 a single deployment, it is unlikely to supplant more technically sophisticated survey tools
378 like ROVs and HOVs. Our study suggests that it is not a good survey tool for cryptic
379 fishes like cabezon, although this may be typical for stationary underwater video

380 platforms used in complex habitat (Watson et al. 2005). Additionally, it is unknown how
381 our use of artificial lighting may have influenced the numbers of fish of different species
382 that we observed, although some prior studies suggest that lighting from stationary HOVs
383 does not seem to have observable effects on many common demersal species found in the
384 U.S. Pacific northwest, including many rockfishes (Hixon et al. 1991, Stein et al. 1992).
385 Despite these limitations, our video lander system may have some advantages for specific
386 sampling situations. Our type of system might be very useful for surveying marine
387 habitats that are risky for larger ROVs or HOVs due to shallow depth, extreme
388 topography or strong currents. We successfully sampled reef systems that are nearshore,
389 but also beyond the depth ranges that are most amenable to scuba surveys. In these
390 habitats, we encountered highly variable visibility, a factor that may typically complicate
391 surveys of these habitats in marine waters of the U.S. Pacific northwest. The smaller
392 vessels that can deploy the video lander, along with its simplicity and portability make
393 mobilizing a survey quick and inexpensive in comparison with most ROVs and HOVs.
394 This should facilitate attempting visual surveys of rocky habitat across a wider range of
395 marine conditions (different days, tidal stages, etc.), increasing the likelihood of
396 encountering favorable conditions and completing a survey in the nearshore environment.
397 This ease of deployment also makes the video lander an excellent choice for very broad-
398 scale surveys of demersal fish distribution, benthic habitat types or species-habitat
399 associations: situations in which broad spatial coverage may be more important than
400 viewing a large amount of seafloor in a single deployment, as is more typical for ROVs
401 and HOVs.
402

403 The reliable retrievability of the video lander frame system suggests the possibility of
404 adapting it for other visual sampling tools. It can be modified to carry other types of
405 equipment, such as still cameras or high-definition video systems or different types of
406 lights. The potential also exists for equipping the lander frame with a calibrated stereo-
407 video or stereo-camera system (e.g. Williams et al. 2010) to estimate the lengths of the
408 fish observed. For the reefs we surveyed, it's also possible that the efficiency of visual
409 data collection with a video lander could be improved for some species of interest (e.g
410 yelloweye rockfish) by using bait as an attractant, an approach that has been shown to
411 very effective in other areas for some species (Willis and Babcock 2000, Watson et al.
412 2005, Harvey et al. 2007, Langlois et al. 2010). This is a topic that warrants additional
413 investigation.

414

415 **5.0 Acknowledgements**

416 Duane Edwards provided the commercial fishing vessel *Maggie*, Craig Taunton and Dick
417 Murray provided the commercial passenger fishing vessel (CPFV) *Endeavor*, and Mike
418 Sorenson provided the CPFV *Miss Raven* as sampling platforms. Field sampling was
419 assisted by Polly Rankin, Steve Jones, Scott Malvitch, Josie Thompson, Ryan Easton and
420 Aaron Chappell. Funding was provided in part by an Oregon Department of Fish and
421 Wildlife grant from the U.S. Fish and Wildlife Service State Wildlife Grant program
422 (grants #OR T-22-C-1 N-05 and #OR T-30-C N2-01).

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515 Figure Captions

516 Figure 1. The video lander, showing the location of sequential break-away connections
517 (arrows), camera (A), lights (B), pressure housing containing batteries and
518 camcorder (C), sacrificial (breakaway) base (D) and steel weight (E).

519

520 Figure 2. Map showing the 5 reef complexes surveyed in initial field tests of the video
521 lander, October 2009 through September 2010.

522

523 Figure 3. Sampling stations (small dots) and the number of yelloweye rockfish
524 (maximum number in a single videof rame) observed (fish symbols) during the
525 2011 survey of the northern portion of Stonewall Bank.

526

527 Figure 4. Still image from video lander footage at Stonewall Bank, Oregon, showing a
528 view that is partially obstructed on the right by nearby rock with a yellowtail
529 rockfish (*Sebastes flavidus*) visible in the left foreground and a yelloweye
530 rockfish visible below it in association with crevice habitat.

531

532 Figure 5. Percent composition of primary and secondary habitat types (see text)
533 identified at stations inside and outside of the Yelloweye Rockfish Conservation
534 Area at Stonewall Bank (Figure 3), Oregon, April through August, 2011.

535

536 Figure 6. Spatial distribution of habitat types preferred by yelloweye rockfish (crevice as
537 a primary habitat type and large boulder as a secondary habitat type) at Stonewall
538 Bank, Oregon, April through August 2011. One possible alternative western

539 boundary for the Yelloweye Rockfish Conservation Area is also shown (heavy
540 hatched line).

Table 1. Design, manufacturer and approximate cost (U.S. dollars) for a video lander camera system with four sacrificial bases.

Component	Design or model number	Manufacturer	Number	Total Cost
Lander frame	32 mm OD aluminum tubing	Local manufacture	1	\$620
Sacrificial base	13 mm diameter mild steel rod	Local manufacture	4	\$450
Base weights	51 mm diameter mild steel rod, 61 cm long	Local manufacture	4	\$47
Camera	Multi-SeaCam [®] 2060	Deep Sea Power and Light	1	\$2,250
Lights	LED Mini-Sealite [®]	Deep Sea Power and Light	2	\$2,600
Pressure housing	102 mm ID aluminum, anodized	The Sexton Company LLC	1	\$1,714
Batteries	13.2 volt, 4.0 Amp-h X 2, NiMH	Energy Sales	4	\$712
Battery chargers	CCC-3029, temperature controlled	Energy Sales	2	\$450
Camcorder	TRV11	Sony	1	\$599
Pressure-activated switch	PSW-3	The Sexton Company LLC	1	\$380
LANC controller board	Micro controller circuit board	Mac Marine Instruments	1	\$250
Underwater cable and connectors	-	Teledyne Impulse	5	\$702
Total				\$10,774

Table 2. Criteria used to classify visibility and the quality of the view from video lander footage of Oregon rocky reef survey sites, October 2009 to August 2011.

Category	Class	Description
Visibility	Poor	View of surrounding substrate completely obscured by turbidity or marine snow
	Moderate	View of surrounding substrate is not obscured but viewing distance is limited by variable turbidity and /or marine snow
	Good	View of surrounding substrate is clear to the limit of the lighted area
View quality	Poor	Camera is facing a rock in close proximity, looking straight down at the substrate or up into open water
	Moderate	View is at acceptable angle, substrate can be seen, however a portion of the view is blocked by nearby habitat
	Good	View is at an acceptable angle and view is unobstructed.

Table 3. Criteria used to classify primary and secondary habitat types viewed with the video lander at the rocky reef survey sites, Oregon, October 2009 through August 2011.

Primary and secondary habitat types	Description
Flat bedrock	Rock with little to no relief
Bedrock outcrop	Solid rock with some relief extending across the view
Large boulder	Boulders approximately 1-3 m in diameter (includes angular blocks of broken bedrock)
Small boulder	Boulders approximately 0.25-1 m diameter
Cobble	Cobble approximately 6-25 cm diameter
Gravel/pebble	Gravel or pebble approximately 2-60 mm diameter
Sand	Sand with grain size 0.06-2 mm diameter
Crevice	Crevices in rock up to 1 m high by 1-3 m wide
Vertical wall	Rock wall higher than 2 m and greater than 80 degrees to the horizontal
Hash	Small broken bits of shell

Table 4. Total numbers (percentage) of fish observed (sum of maximum counts across stations for the 16 most abundant species or groups) in selected visual fish surveys with an experimental video lander, by species or group and survey area, October 2009 through September 2010 (n denotes number of stations sampled).

Fish species	Cape Perpetua reef (February, n=30)	Cape Perpetua reef (July, n=30)	Seal Rocks reef (n=43)	East Siletz reef (n=36)	West Siletz reef (n=30)	Stonewall Bank (n=173)	Total
Rockfish							
Black rockfish <i>Sebastes melanops</i>	62 (15.4)	72 (21.1)	182 (48.3)	15 (5.5)	11 (6.7)	0 (0.0)	342 (15.1)
Blue rockfish <i>S. mystinus</i>	2 (0.5)	1 (0.3)	18 (4.8)	187 (68.3)	47 (28.5)	54 (7.6)	309 (13.6)
Brown rockfish <i>S. auriculatus</i>	3 (0.8)	1 (0.3)	0 (0.0)	0 (0.0)	0 (0.0)	0 (0.0)	4 (0.2)
Canary rockfish <i>S. pinniger</i>	173 (43.0)	156 (45.8)	47 (12.5)	18 (6.6)	74 (44.9)	202 (28.4)	670 (29.5)
Copper rockfish <i>S. caurinus</i>	6 (1.5)	2 (0.6)	5 (1.3)	0 (0.0)	0 (0.0)	0 (0.0)	13 (0.6)
Unidentified juvenile	7 (1.7)	47 (13.8)	56 (14.9)	3 (1.1)	0 (0.0)	281 (39.5)	394 (17.4)

Pile perch <i>Damalichthys</i>	68 (16.9)	11 (3.2)	6 (1.6)	0 (0.0)	2 (1.2)	0 (0.0)	87 (3.8)
<i>vacca</i>							
Unidentified sculpin Cottidae	1 (0.3)	1 (0.3)	0 (0.0)	2 (0.7)	0 (0.0)	4 (0.6)	8 (0.4)
Total	402	341	377	274	165	711	2,270
Mean station depth (m)	51.3	51.3	30.7	33.6	40.4	54.3	

Table 5. Summary of mean number of fish observed per station for five demersal fish species surveyed with a video lander at Stonewall Bank, inside and outside (to the west) of the Yelloweye Rockfish Conservation Area, April through August 2011 (Figure 3). Also shown is the percentage difference in abundance outside the YRCA and the P-value from a Wilcoxon test comparing mean abundance inside and outside the YRCA.

Species	Mean number (1SE) observed per station		Percentage difference outside YRCA	P-value in Wilcoxon test
	Inside YRCA	Outside YRCA		
Canary rockfish	0.561 (0.140)	0.631 (0.140)	+12.5	0.320
Kelp greenling	0.144 (0.022)	0.114 (0.020)	-20.8	0.303
Lingcod	0.148 (0.028)	0.072 (0.020)	-51.4	0.007
Rosethorn rockfish	0.318 (0.046)	0.228 (0.040)	-28.3	0.093
Yelloweye rockfish	0.144 (0.028)	0.179 (0.032)	+24.3	0.417

Fig. 1

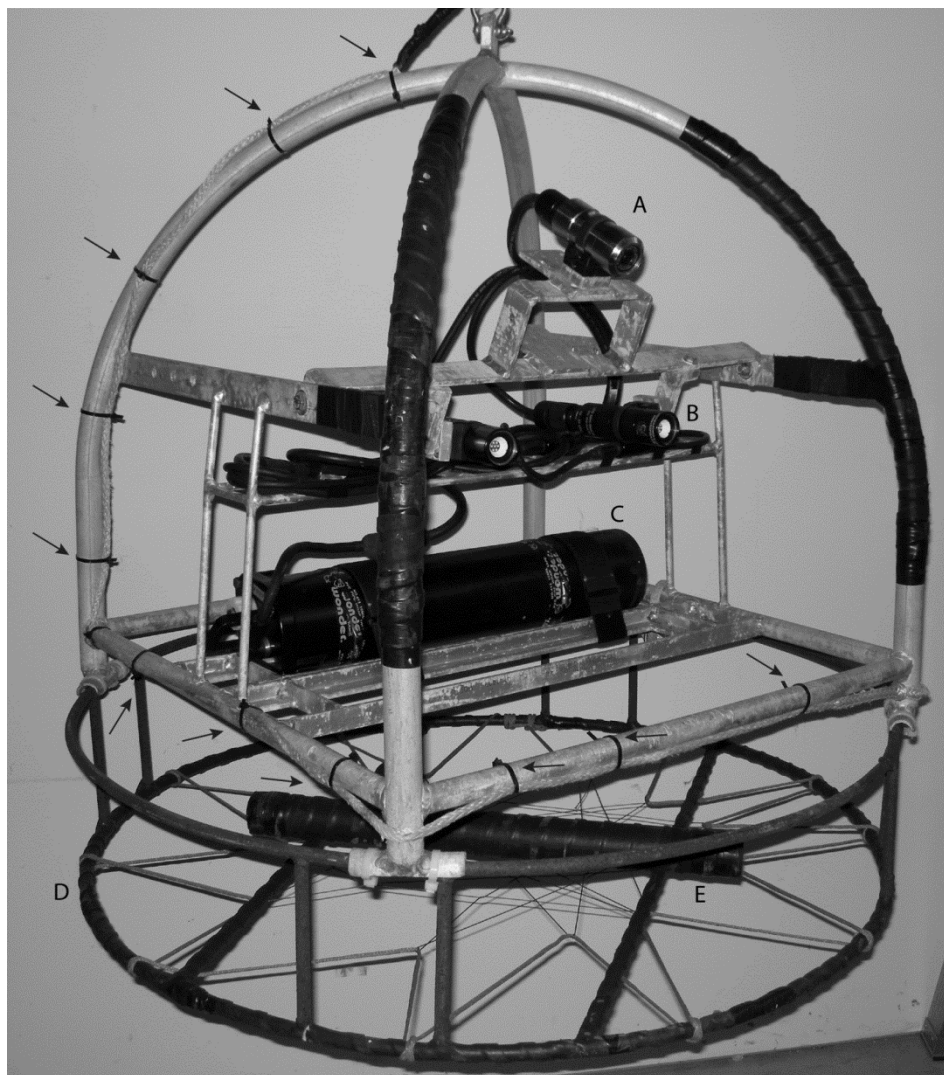


Fig. 2

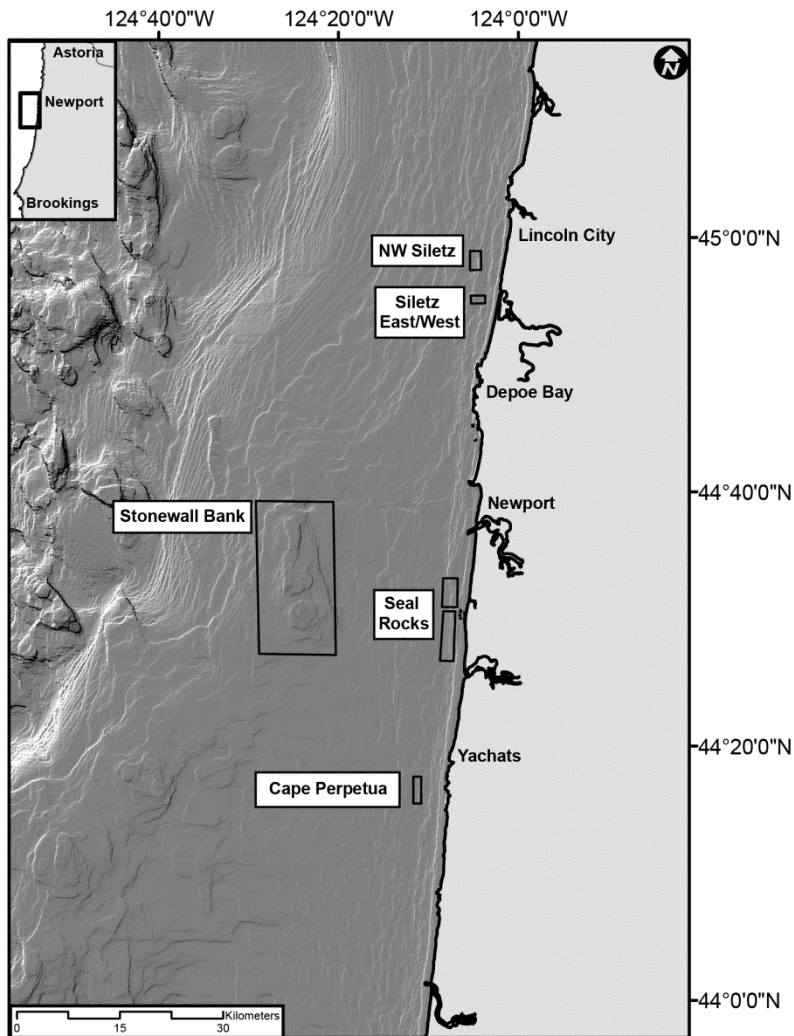


Fig. 3

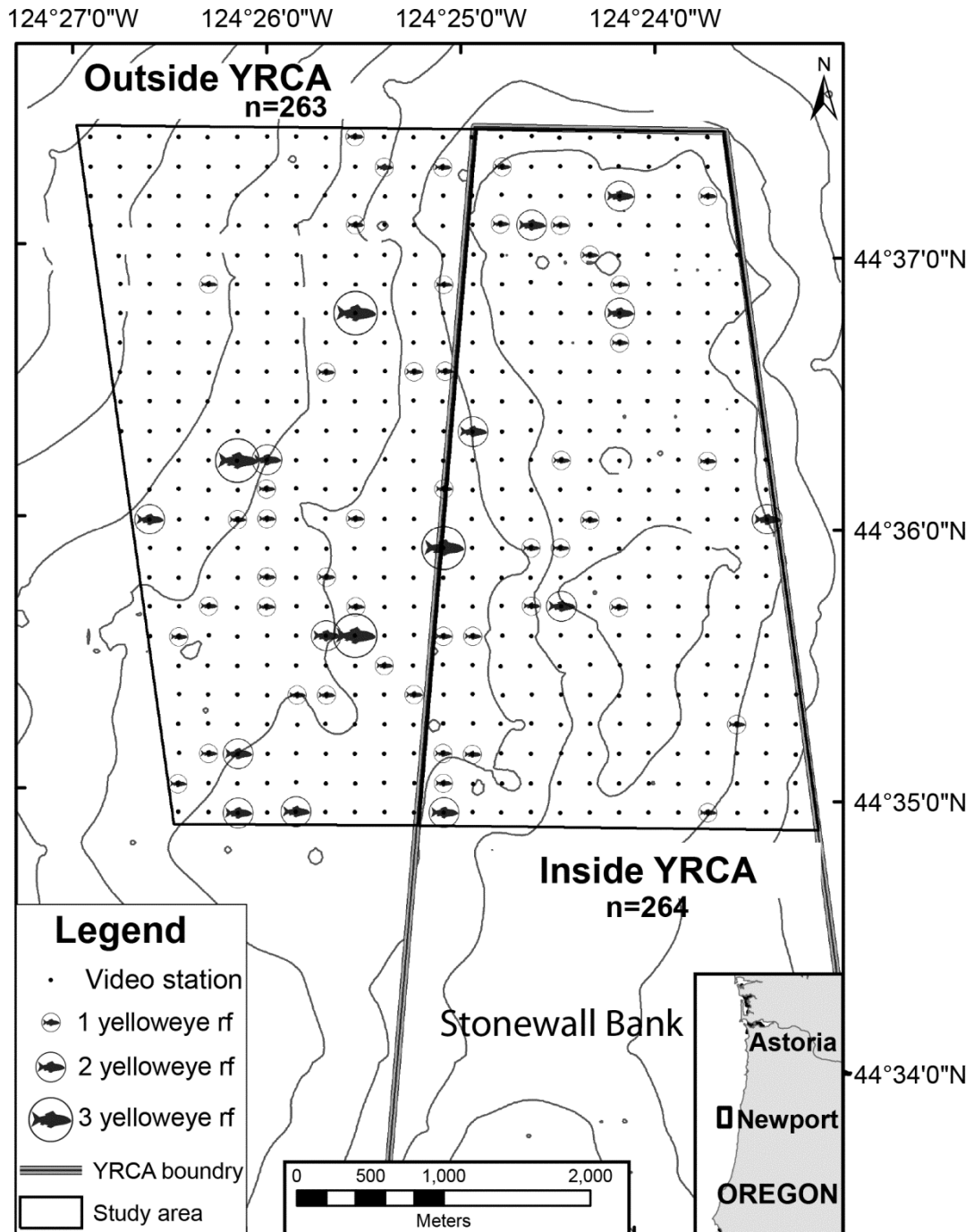


Fig. 5

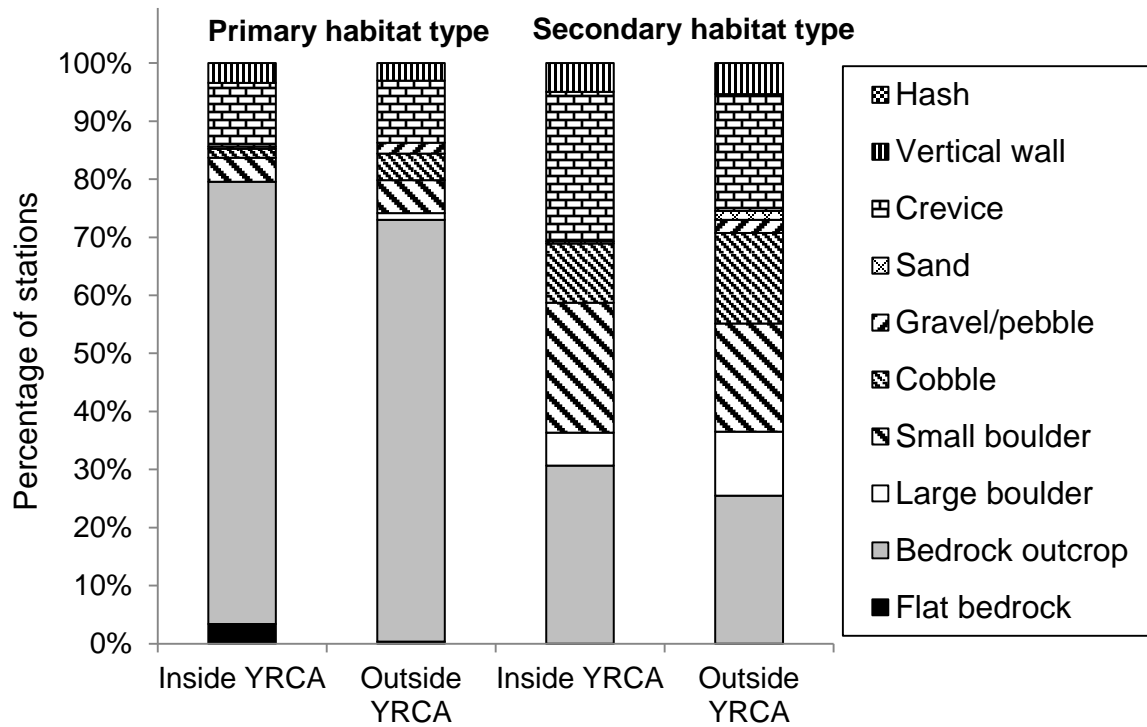


Fig. 6

