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Biogeographically distinct controls on C₃ and C₄ grass distributions: merging community and physiological ecology

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Citation	Griffith, D. M., Anderson, T. M., Osborne, C. P., Strömberg, C. A. E., Forrestel, E. J., & Still, C. J. (2015). Biogeographically distinct controls on C₃ and C₄ grass distributions: merging community and physiological ecology. Global Ecology and Biogeography, 24(3), 304–313. doi:10.1111/geb.12265
DOI	10.1111/geb.12265
Publisher	John Wiley & Sons Ltd.
Version	Accepted Manuscript
Terms of Use	http://cdss.library.oregonstate.edu/sa-termsofuse



- 1 Article Title: Biogeographically distinct controls on C₃ and C₄ grass distributions: merging
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- 16 Short running title (45): Climate disequilibrium in C₄ grass distributions
- 17 Keywords: Biogeography, C₃, C₄, crossover temperature, tree cover, invasive, fire
- 18 Type: **Research Paper**
- Number of words in the abstract including key words (10): **300** of 300
- 20 Main text, including Biosketch (32): **5632** of 5000
- Number of references: **50** of 50
- Number of figures: 4 of 6
- 23 Tables: **1**

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ABSTRACT

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34 Aim C₄ photosynthesis is an adaptation that maintains efficient carbon assimilation in warm and 35 low-CO₂ conditions. Due to the importance of C₄ grasses for carbon and surface energy fluxes 36 numerous models have been proposed to describe their spatial distribution and forecast responses 37 to climate change. These models often rely on broad climate predictors (e.g., temperature and 38 precipitation) but fail to integrate other ecologically relevant factors, such as disturbance and 39 competition, which may modify realized C₃/C₄ grass distributions. We evaluate the contribution 40 of ecological factors, in addition to climate predictors, to C₃/C₄ grass distributions across 41 multiple biogeographic regions of North America in a multi-source database of >40,000 42 vegetation plots. Location Conterminous United States of America (USA). 43 44 **Methods** We identified a comprehensive pool of physiological-climatic models in the literature 45 and used information theoretic criteria to select a primary physiological predictor of C₃ and C₄ 46 grasses. Subsequently, the climate model was combined with ecological predictors using a 47 multiple regression framework and tested within eight regions within the USA. 48 **Results** Surprisingly, grass-dominated communities across the USA exist largely in a C_3 or C_4 49 dominated state. Transitions between C₃/C₄ dominance were best explained by models that 50 integrated temperature and precipitation with ecological factors that varied according to region. 51 For some regions, such as Eastern Temperate Forests, local, ecological factors were comparable 52 in strength to broad climate predictors of C_3/C_4 abundance. 53 **Main conclusion** Local, ecological factors modify C₃/C₄ grass responses to broad-scale climatic 54 drivers in ways that manifest at regional scales. In Eastern Temperate Forests, for example, C₄ 55 grass abundances are maintained below climatic expectations where tree cover creates light

- 56 limitation, but above expectations where frequent fires reduce tree cover. Thus, local ecological
- factors contribute to major among-region differences in the climate responses of C_3/C_4 grasses.

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Main Text:

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INTRODUCTION

As humans continue to modify ecosystems and alter the Earth's climate, predicting future species' distributions and shifting range limits has become a paramount goal in ecology. Many contemporary methods for modeling species distributions (e.g., climate envelope, process-based) are largely based on the assumption that species' physiological tolerances to environmental variation (e.g. temperature, precipitation, etc.) determine species occupancy across the landscape (Merow et al., 2013). However, an alternative perspective argues that local ecological processes (e.g. disturbance, predation, facilitation, storage effects) interact with species' ecophysiological and life history traits to determine species distributions (e.g., Callaway, 1995; Weiher & Keddy, 1995; Araújo & Pearson, 2005; Maestre et al., 2009). Approaches that merge species distribution modeling based on physiological tolerances to variation in climate with community ecological theory (e.g., Guisan & Rahbek, 2011; Scheiter et al., 2013) show tremendous promise in predicting species distributions, as well as assessing the contributions of relevant abiotic and biotic drivers (Araújo et al., 2013). However, integrative methods are not commonly applied in contemporary modeling approaches, especially at large scales (Araújo & Rozenfeld, 2014). The Poaceae (grasses) are a cosmopolitan family of flowering plants that dominate the herbaceous layer of several major biomes which together cover as much as half the Earth's land surface (grasslands, savannas and managed rangelands - Asner et al., 2004). A central factor determining the primary production of grass-dominated ecosystems is the proportion of species that fix carbon using the C₄ photosynthetic pathway. This is because C₄ species have evolved an efficient mechanism for reducing the negative effects of photorespiration in warm climates

compared to C₃ species (Ehleringer *et al.*, 1997), creating differences in carbon capture, water use, phenology, and quantum yield (Still *et al.*, 2003a,b; Randerson *et al.*, 2005; Taylor *et al.*, 2010, 2014). In contrast, C₃ grasses have a competitive advantage in cool environments where photorespiration is reduced (Sage, 2004). As such, C₃ and C₄ grasses are often observed to segregate along temperature gradients; historically, broad-scale climate predictors such as temperature and precipitation have been seen as the predominant determinants of C₃/C₄ grass distribution. While, local, and often stochastic, factors like fire, herbivory, and competition (in addition to temperature) are known to modify the C₃:C₄ grass ratio (e.g., Heckathorn *et al.*, 1999), these various drivers have not been incorporated into a single framework for predicting C₃/C₄ grass distributions. Thus, the history of climate-based research combined with community ecological research on C₃ and C₄ grasses provides an ideal system for testing models that merge physiological and ecological processes in order to predict the abundance and distribution of these functional types.

Several studies have investigated the distribution of C₃ and C₄ grasses along temperature gradients across a range of spatial scales. Teeri & Stowe (1976) first showed that the percentage of C₄ grass species in regional floras was closely tied to the minimum July temperature and the length of the frost-free period, suggesting that cold growing season temperatures limit C₄ grasses (while warm temperatures favor them). Numerous other studies have confirmed a role for temperature in C₄ grass distributions (reviewed in Ehleringer *et al.*, 1997; Sage *et al.*, 1999). However, additional factors such as soil texture (Epstein *et al.*, 1997) and growing season precipitation (Paruelo & Lauenroth, 1996; Winslow *et al.*, 2003; Murphy & Bowman, 2007; von

Fischer *et al.*, 2008) often add considerably to explained variance and to predicting the spatial distribution of C₃ or C₄ grasses.

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The prevailing physiological hypothesis for explaining temperature-driven patterns of grass functional-type distributions is that C₃ and C₄ species differ in the temperature sensitivity of their quantum yield (the amount of carbon fixed per unit of light absorbed; Ehleringer et al., 1997). At higher temperatures C_3 quantum yield is reduced by photorespiration while at lower temperatures the additional energetic requirements of the C₄ carbon concentrating mechanism generally result in lower quantum yield for C₄ grasses (Ehleringer, 1978). This model implies the existence of a 'crossover temperature' at which the photosynthetic gains of C₃ and C₄ grasses are equal; above this temperature, C₄ grasses have higher photosynthetic and growth rates, whereas the opposite holds below this crossover temperature. For global-scale distribution and carbon models, crossover temperatures have been combined with assumptions about the minimum quantity of precipitation necessary for grass growth to predict the climates in which C₄ or C₃ grasses should dominate. The resulting prediction is that in months with mean air temperature ≥ 22 °C and rainfall >25 mm C₄ should out-compete C₃ grasses (Collatz et al., 1998; Still et al., 2003a). However, studies which take a phylogenetic approach have found that air temperature does not always explain the distributions of closely related C₄ and C₃ species, implying that thermal adaptations in grasses may be indirectly associated with photosynthetic functional type (Edwards & Still, 2008; Pau et al., 2013; Still et al., 2013).

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Where C_4 and C_3 grasses coexist, C_3 grasses may gain a competitive advantage if they can acquire resources (e.g. space, nutrients) early in the growing season before C_4 plants become

most active (Ode *et al.*, 1980; Barnes *et al.*, 1983; Tieszen *et al.*, 1997). Conversely, locations with cool spring/fall seasons that should favor C₃ grasses may be completely dominated by C₄ grasses which preempted resources during the warm summer season (Tieszen *et al.*, 1997). These seasonal competitive effects have the potential to promote dominance of grasses that are not climatically favored at a given location and time. Similarly, in forested areas, shading from trees also has potential to favor C₃ grasses over C₄ because they generally have a carbon gain advantage under low light conditions (Sage *et al.*, 1999; Peterson *et al.*, 2007).

Several other traits related to photosynthetic pathway alter the competitive interactions between C₃ and C₄ grasses, such as how species respond to disturbances such as herbivory and fire (Monson et al., 1983; Heckathorn et al., 1999). Compared to C₃ grasses, C₄ species are often better equipped for rapid regrowth in the high-light conditions following defoliation (Heckathorn et al., 1999). Furthermore, C₄ grasses in savannas are believed to express traits that promote understory fires, such as leaf flammability, which then maintain high light environments by reducing tree cover in savannas (Ratnam et al., 2011; Veldman et al., 2013). From the perspective of community ecologists, the mechanisms promoting coexistence or dominance of C₃ and C₄ species at individual sites are best explained by temporal or spatial resource partitioning by species with different resource requirements (Monson et al., 1983; Tilman & Pacala, 1993; Fargione & Tilman, 2005). Finally, species and lineages are likely to have distinct and independent responses to temperature, precipitation, and other environmental factors (regardless of functional type) that we expect to contribute to the variation in grass distributions (Edwards & Still, 2008).

While the broad-scale distribution of C₃ and C₄ grasses has been studied extensively in North America, previous studies have been limited because they (i) relied upon indirect proxies of relative abundance (e.g., floral survey lists, soil organic carbon stable isotopes and land cover classifications), (ii) have been limited in spatial extent, and (iii) focused largely on climatic and physiological explanations (Teeri & Stowe, 1976; Paruelo & Lauenroth, 1996; Epstein *et al.*, 1997; Tieszen *et al.*, 1997; Sage *et al.*, 1999). Using plot level abundance data, we set out to explore the degree to which direct measurements of vegetation abundance support hypothesized models of C₃/C₄ distributions, including those used to estimate global carbon budgets.

Specifically, we intend to identify if ecological mechanisms not directly related to the efficiency of photosynthesis produce disequilibrium with climate that is discernible in species distribution models at broad scales. Finally, we ask whether integrating a selection of ecological predictors can increase the explanatory power of the climate model and provide support for specific local mechanisms.

METHODS

Vegetation plot data

Vegetation plots with cover abundance measurements were sourced from databases, literature sources, and unpublished sources (Appendix S1 in Supporting Information). Plot data met the following criteria: (1) sample areas were between 100 and 1000 m^2 , (2) accurate spatial data were provided, (3) plant abundance in the herbaceous layer was collected by species, and (4) plots contained species from the Poaceae (mean grass abundance was 65 %). Criterion (4) was included because our primary goal was to understand the controls over the ratio of C_3 : C_4 grasses, not those factors that determine the abundance of grass relative to other plant taxa (herbaceous forbs, shrubs, etc.). Consequently, our response variable, proportion of C_4 grass, was calculated

for each plot as the sum of C₄ grass abundance divided by the total grass abundance. Abundance was measured as aerial estimates of percent cover, but where cover classification systems (e.g. Carolina Vegetation Survey cover classes) were used, we converted cover ranges to the midpoint percent cover. Photosynthetic functional types were assigned to species using (Osborne *et al.*, 2014).

Explanatory variables

We reconstructed several previously published statistical models of C₃/C₄ grass distributions using climate 30 year (1971-2000) climate normals for the USA, sourced from the PRISM Climate Group (http://www.prism.oregonstate.edu/; details in Appendix S1). These are: (1) Teeri & Stowe's (1976) July minimum temperature and consecutive frost-free months predictors, (2) Paruelo & Lauenroth's (1996) growing season precipitation model, the (3) Epstein *et al.* (1997) soil texture and climate model, the (4) July temperature and rainfall model of von Fischer *et al.* (2008), and (5) the Collatz *et al.* (1998) crossover temperature model. The construction of the crossover model is describe below as it is the predominant model in the literature and requires some additional explanation.

The model representing the C₃/C₄ crossover temperature was created by applying a set of climatic criteria (temperature and rainfall thresholds) to all grid cells in the monthly climate dataset, following the work of Collatz et al. (1998) and Still et al. (2003a). Because multiple crossover temperature values have been reported in the literature, we created separate models spanning the entire range of empirical temperature thresholds (5-31 °C) separately for minimum, mean, and maximum monthly temperatures (Ehleringer *et al.*, 1997). Within each grid cell, each

month with sufficient simultaneous rainfall (\geq 25 mm) and temperatures above the crossover threshold were classified as favoring the growth of C_4 grasses over C_3 grasses (Still *et al.*, 2003a; Pau *et al.*, 2013; summarized in Fig. S2). We then summed the number of months in each grid cell that favored C_4 growth to produce a metric, hereafter referred to as the number of "months favoring C_4 grasses" or the "crossover temperature model", to be used as a predictor of the expected C_4 grass proportion. This model, often called the Collatz model in the literature, always refers to both the temperature and rainfall criteria together. The best minimum, mean, and maximum monthly temperatures for use in the crossover temperature model was selected by comparing their variance explanations when each was regressed against the proportion C_4 grasses in the plot as a response variable.

In addition to minimum and maximum temperatures, mean annual temperature (MAT), mean annual precipitation (MAP), the crossover temperature model, frost-free months, and seasonal rainfall, we also extracted non-climatic predictors such as tree cover, soils (cation exchange capacity, organic carbon, pH, and texture), fire frequency, and proportion invasive grasses (Appendix S1). We included invasive grasses as a predictor because of the observation that exotic species may possess different relationships with disturbance regimes or otherwise interact with native species in a way that increases their representation in the community (D'Antonio & Vitousek, 1992; Smith & Knapp, 1999).

Data analysis

Data analyses were aimed at (1) determining the best possible model that predicts C_3/C_4 grass abundance based of physiological limits to temperate and precipitation (i.e. climate variables)

and (2) to what extent the predictions of the climate model were modified by local ecological factors (e.g., fire, competition, see introduction) at regional scales. In the first step, an appropriate analytical model of grass physiology was chosen using formal model selection procedures based on the Akaike Information Criterion (AIC; Appendix S1) to compare the support among previously published C₃/C₄ distribution models (e.g., Teeri & Stowe, 1976; Paruelo & Lauenroth, 1996; Epstein *et al.*, 1997; Still *et al.*, 2003a; von Fischer *et al.*, 2008) across the entire dataset. Then, we inspected the fit of the physiological-climatic C₃/C₄ distribution model to the observed vegetation data from across the USA. To control for phylogenetic, biogeographic, and historical differences (e.g., history of competition, disturbance) between regions of North America, we partitioned our data into geographical subsets based on the regions described by Omernik (see Appendix S2).

In the second step, we developed a series of "verbal" models to represent various potential interactions between climate and ecological drivers (Fig. 1) as a framework for exploring the ways that local ecological factors might cause C_3/C_4 grass distributions to be in disequilibrium with climate. Our intention is not to provide an exhaustive set of models or directly infer specific ecological processes from these patterns; rather the goal of this exploratory analysis is to provide a rational for statistically testing for the effects of various ecological predictors in the next step of analysis. Fig. 1A shows our expectation if C_3/C_4 grasses are in climate equilibrium— C_4 species exist when and where they are favored and are absent when conditions are never physiologically suitable (Araújo & Pearson, 2005). The remaining "verbal" models (Fig. 1B-F) illustrate patterns that can be expected when other mechanisms (interspecific competition, disturbance feedbacks, and species interactions) play important roles in addition to the purely physiological model. For

example, if C₃ grasses are able to persist in a grassland, perhaps because they can capture and store resources during the cool season, then C₄ species may never attain the level of dominance predicted by the local climate until C₃ grasses are incapable of growing (Fig. 1B). On the other hand, disturbance favoring C₄ grass species (e.g., fire) might maintain dominance of C₄ grasses at sites where the climate favors C₃ grasses (Fig. 1C). Competition and disturbance are just two straight-forward examples that would produce disequilibrium with climate. Furthermore, multiple factors might operate in conjunction—for instance, cool season competition from C₃ grasses could prevent establishment of C₄ species in low temperatures environments, but in warmer conditions C₄ grasses might be promoted by fire (Fig. 1D). Another possibility is that C₄ species do not grow in areas where temperatures favor them for only short periods (e.g., one month) because there is not sufficient time to establish a population (Fig 1E). Finally, in warm environments where C₄ grasses are expected to dominate exclusively, C₃ species may continue to persist if they store resources gained during temporarily beneficial environmental conditions, such as following a frost event that kills competing C₄ grasses (Fig. 1F; "storage effect"; Chesson, 2000).

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Treating each ecoregion separately, we compared the physiological model expectations (e.g., Fig. 1A) to the observed plot-based C₄ grass proportions with the goal of identifying patterns of deviation indicative disequilibrium processes (e.g., Fig 1B-F, described above). For each ecoregion containing vegetation data, we statistically tested whether the observed versus expected relationship reflected climatic equilibrium (i.e., a linear fit) or disequilibrium (i.e., nonlinear, requiring a second or third degree polynomial term in a regression model). Linearity is a good representation of climate equilibrium because it indicates that as climate changes there

are corresponding changes in C_4 grass abundances. In addition to assessing linearity, we tested the hypothesis that plots existed largely in either low or high C_4 states by applying Hartigan's dip test to test for and quantify the degree of multimodality in each ecoregion (Maechler 2013).

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Finally, in ecoregions in which grass distributions were likely to be in disequilibrium with climate, we assessed the degree of statistical support for a selection of other potential mechanisms that could produce the observed patterns. We did this by fitting models which test for partial effects of additional predictors on the proportion of C₄ grasses after the climate-related predictor variable was accounted for. These additional predictors were: (1) fire, (2) tree cover, (3) proportion of grasses that are invasive, and (4) soil characteristics (Appendix S1). While temperature and rainfall are already included in the Collatz crossover model, min and max temperature and rainfall were included in our regressions because we considered that additional temperature thresholds (tolerance to extremes) and interactions with rainfall might operate in addition to the crossover model and be important in particular regions (Still et al., 2013). Due to the spatial autocorrelation and bounded nature of our proportional plot data we fit models using boosted beta regression (Appendix S1). This approach is appropriate for modeling spatial data with beta distributed responses (i.e. bounded zero to one) when accounting for partial effects and modeling how variance responds to predictors. For example, our regressions modeled variance through phi (ϕ) , which is a parameter of beta distributions that describes the degree to which a variable (in this case, proportion of C_4 grass) is either hump-shaped (large ϕ) or U-shaped (small φ)—this shape can change in response to predictor variables. In this case a positive effect of a predictor on ϕ would indicate that variance decreases around the mean response; conversely, variance increases as ϕ decreases.

There are several aspects of our analytical approach that deserve a caveat. First, the sampling dates for our plot data were skewed towards summer months and to reduce bias we restricted our analysis to summer month to represent the height of the growing season; this means we may have missed patterns that might only be evident when all seasons are fully sampled. Second, in this study we have lumped species into functional type even though these species will have individualistic responses to the environment that could explain additional variation or be confounded with our other predictors. Lastly, our modeling approach mostly considers main effects of temperature and ecological predictors even though temperature likely interacts with biotic factors to produce ecological outcomes (Dillon *et al.*, 2009).

RESULTS

The number of months that favor C_4 grasses (i.e. the Collatz crossover model) based on a monthly maximum of 27 °C and a minimum 25 mm rainfall emerged as the best physiological predictor of the observed C_4 grass proportion (R^2 = 0.40; Fig. 2); the max 27 °C Collatz model had the lowest AIC (and highest R^2) of all single predictor models from the literature (Table S3) and had a strong positive effect in all ecoregions (Tables 1 and S3). The quality and spatial coverage of the plot data allowed us, for the first time, to empirically derive the best crossover temperature for the Collatz model (Fig. 2) and to compare it to other models with a modern statistical approach. The next best models were 9 °C and 18 °C for the minimum and mean temperature Collatz crossover models, respectively. Consequently, in downstream analyses the crossover temperature model was used to represent the best physiological-climate predictor of the proportion of C_4 grasses.

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When the proportion of C₄ grasses in a plot was regressed upon the number of months favoring C₄ growth, all of the ecoregions, with the exception of the Mediterranean, showed a non-linear relationship best fit by either a second or third degree polynomial (Table S3). To better visualize the middle range and upper boundary of these responses we used quantile regression to model the median and upper 95th quantile of C₄ proportion; because the relationships were non-linear we fit these models with additive components (Fig 3). The modeled upper limit of C₄ grasses was consistently above 50% C₄ in all regions except for the Mediterranean regions. In the Great Plains, Temperate Sierras, and Southern Semiarid Highlands, the median C₄ response is higher than expected when just one month favors C₄ grasses. The Eastern Temperate Forests and North American Deserts have sigmoidal relationships between predicted and actual C₄ proportion, and their median curves (and the Northern Forest and Northwestern Forested Mountain curves) remain below the physiology-climate predictions even when several months favor C₄ (Fig. 3). One particularly compelling results was that the distribution of C₄ grass proportions in all regions (except Mediterranean) could not be explained by a unimodal distribution (for all regions: D=0.01-0.13 and p < 0.001 from Hartigan's dip test) —i.e., they were at least bimodal with peaks near 0 and 1 (Fig. 3). The Mediterranean showed a linear pattern and was unimodal which might reflect C₃ competitive dominance or simply that C₄ is only favored for up to two months. These results establish that seven of eight ecoregions showed nonlinear relationships between observed and expected C₄ proportion. Because we found significant deviations from our predictions, we used boosted beta regression in order to determine if fire, tree cover, invasiveness, soils, min and max temperature, or rainfall could explain additional variation in C₄

grass proportion after accounting for physiological-climate effects based on the crossover temperature model (with 27 °C temperature and 25 mm rainfall criteria). The explanatory power of all models was increased by the addition of these variables (Table 1; Appendix S1), compared to models with only the number of months favoring C₄ grasses as the predictor. R² values increased from 0.18 to 0.30 in Eastern Temperate Forests, 0.19 to 0.41 in Great Plains, 0.33 to 0.51 in North American Deserts, 0.23 to 0.42 in Temperate Sierras, and from 0.49 to 0.53 in Southern Semiarid Highlands.

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In Eastern Temperate Forests, there was a strong negative effect of tree cover on C₄ grass proportions, whereas invasive grass proportion, soil organic carbon, and minimum temperature each had positive effects (Table 1). Fire did not have a main effect but decreased the variance associated with C_4 (ϕ coefficient = 0.06; see Methods) meaning that where fire is common it is rare to find C₃ grass species. In the Great Plains, the crossover temperature model was the only significant predictor of the mean response, but, tree cover (ϕ coefficient = -0.009) and increasing minimum temperatures (ϕ coefficient = -0.005) tended to increase variance there. In contrast, increasing maximum temperatures were positively associated with C₄ grasses in the Deserts, and both minimum and maximum temperatures were important in the Temperate Sierras. MAP was associated with decreasing C₄ grass presence in both Temperate Sierras and the Semiarid Highlands. Unexplained spatial patterns were present in the Great Plains, Temperate Sierras, and North American Deserts datasets as demonstrated by the significant fit of geographic coordinates to the data in our boosted beta regressions (Table 1; Appendix S1). Finally, in three regions where invasiveness was important, an associated decreased in variance of proportion of C₄ grass was observed.

DISCUSSION

To our knowledge this is the first time that the C₄ grass fraction from plot abundance measurements of specific taxa (rather than floral lists, aggregated presence/absence, or herbarium collections data) has been compared quantitatively at a broad scale to the state-of-theart models of C₄ distribution (Teeri & Stowe, 1976; Sage *et al.*, 1999; Murphy & Bowman, 2007; Still *et al.*, 2013). In doing so, we revealed that there are major, biogeographically distinct deviations from climate equilibrium that exist in the broad scale spatial distributions of two highly studied grass functional types.

Based on previous work (Still *et al.*, 2003a), we expected to find that the physiological model using a mean 22 °C crossover temperature criteria (and 25 mm rainfall screen) for counting the number of months favoring C_4 grasses would best account for the variation in the data. Instead, we found that a maximum temperature of 27 °C was a better predictor of C_4 grass abundance than all other models (Fig. 1 and Table S3). One possible explanation for this result is that mean temperatures integrate daytime and nighttime temperatures, while maximum temperature better represents the daytime growing conditions (especially mid-morning when stomata are most open) of grasses and should reflect the C_4 advantage in reducing photorespiratory costs relative to C_3 grasses. Our result is consistent with empirical data suggesting that, when temperature is represented by daytime measurements, it results in high crossover temperatures (Ehleringer *et al.*, 1997; Sage *et al.*, 1999). Although it is not surprising that temperature was a good predictor of C_4 grass distribution, the more significant and novel finding is that the temperature model is

not adequate to explain these distributions on its own. Whereas the number of months favoring C_4 grasses explained the highest percentage of the variance in C_4 grass abundance across the entire dataset (40 %), the residual variation not explained by temperature is considerable and the region-specific analyses show that ecological variables, such as fire and tree cover, play an important role in determining C_3/C_4 grass distributions.

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Seven of our eight regions were characterized by non-linear relationships between the observed and predicted results from the temperature/precipitation model, indicating that there are factors other than climate that are determining grass functional type distributions at large scales (Fig. 3). Moreover, deviations from climate equilibrium were unique to different regions which suggests the presence of region-specific processes or mechanisms that modify grass climate responses in areas with different biogeographic histories. These non-linear patterns suggest different types of factors that might modify climate-determined grass distributions (introduced in Fig. 1), such as seasonal competition favoring either C_3 or C_4 , disturbance factors like fire or herbivory favoring either C₃ or C₄, minimum growing season thresholds, or coexistence mechanisms such as storage effects (Brown & Maurer, 1989; Chesson, 2000; Araújo & Pearson, 2005; Araújo et al., 2013). In two regions, Eastern Temperate Forests and North American Deserts, the median quantile regression results exhibit a sigmoidal-shaped curve implying processes that suppress C₄ grasses in shorter warm growth seasons but promote C₄ above expected beyond a five month warm growing season. Two generally forest-dominated ecoregions, Northern Forests and Northwestern Forested Mountains, have a restricted number of months favoring C₄ grasses because they occupy limited climatic space. Interestingly, these regions have their median responses suppressed below expectations, whereas regions with more open canopy or generally non-forest

habitat tend to approach maximum C₄ proportions rapidly (Southern Semiarid Highlands, Temperate Sierras, and Great Plains but not North American Deserts). Altogether, these patterns point to the existence of biogeographically and ecologically distinct factors and processes influencing the observed distributions of C₄ and C₃ grasses.

In the boosted beta regressions, where we modeled the C₄ proportion of grasses as a function of several covariates, the strength of the relationship and the degree to which other predictors were needed to explain variance in C₄ grass proportion varied among regions. These results support the assertion that North America ecoregions differ in the degree to which temperature and physiology control C₄ distributions. It is also consistent with the hypothesis that the global rise to dominance of C₄ grasses cannot be explained solely by photosynthetic pathway; instead, other adaptations may be equally important to explaining dominance (Edwards *et al.*, 2010). For example, one proposition is that fire-adapted C₄ grasses facilitated the expansion of the grass-dominated savanna systems in the late Miocene (Scheiter *et al.*, 2012). For example, in our dataset the two most dominant species in the southeast were the fire-adapted C₄ species *Aristida stricta* and *A. beyrichiana*, whereas across the rest of the temperate eastern US the dominant grasses were C₃, and likely less fire-prone species *Brachyelytrum erectum* and *Danthonia spicata*. However, the dominant C₄ species still include fire-adapted species like *Andropogon gerardii* and *Schizachyrium scoparium* in addition to the invasive *Microstegium vimineum*.

The only effect of fire regime found in this study was a negative correlation between fire frequency and the variance associated with C_4 proportion in Eastern Temperate Forests. This supports the frequently reported association of C_4 with fire, including in North America, because

at the scale of our analysis fire maintains a high proportion of C₄ grass (D'Antonio & Vitousek, 1992; Scheiter et al., 2012; Veldman et al., 2013). This result introduces the possibility that in the warmer areas of Eastern Temperate Forests (e.g., Florida) high relative abundances of C₄ are already explained by the physiological model, and the influence of fire is mediated through a negative impact on tree cover. A known indirect effect of fire that promotes C₄ grasses is the reduction of forest trees (as opposed to savanna trees) that would otherwise negatively influence the C₄ understory (Veldman et al., 2013). Therefore, fire effects on C₄ might be masked by the negative impact of tree cover, found in three ecoregions. To explore the potential for indirect effects and the influence of scale on our results, we conducted a heuristic analysis comparing the effects of fire on C₄ grass proportion within and outside of the natural range of Longleaf pine (*Pinus palustris*), a species characteristic of the pyrogenic grasslands of the Southeastern United States (see Appendix S3). This analysis showed that C₄ proportion was still bimodal, whether within or out of the range of Longleaf pine. However, C₄ grasses were more common within this range, whereas C₃ grasses were more common in the rest of the Eastern Temperate Forests (Fig. 4; D = 0.06-0.11, p < 0.001). Furthermore, a simple path analysis revealed direct (fire increasing C₄ abundance) and indirect (mediated through reduced tree cover) effects of fire on C₄ grasses inside of the Longleaf pine range but not outside the range (Fig. 4; Appendix S3). This pattern is likely the indirect consequence of shading and the creation of a cool microclimate, relating to the fact C₄ grass species are thought to be at a competitive disadvantage in shade and less able to utilize sunflecks than C₃ grasses (Horton & Neufeld, 1998; Sage et al., 1999). This effect seems to be modified by the presence of the invasive, apparently shade tolerant C₄ grass M. vimineum (Horton & Neufeld, 1998). Accordingly, invasive grass (M. vimineum) abundance had a positive influence on the proportion of C₄ grass in the Eastern Temperate Forests.

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In contrast, grass invasiveness had negative effects on the abundance of C₄ grasses in both North American Deserts and Temperate Sierras, where the C₃ invasive *Bromus tectorum* is responsible for reducing the C₄ proportion to below expected values. The numerous other invasive grasses in the dataset (e.g., *Eragrostis curvula*—C₄ and *Dactylis glomerata*—C₃) are all in low abundance and appear to have less of an influence on C₄ cover. Invasives may shift the proportion of C₄ grasses through diversity reduction and competitively exclusion or through interaction with disturbance (D'Antonio & Vitousek, 1992; Smith & Knapp, 1999). In general, all mean effects of invasive grasses on C₄ proportion (positive and negative) in our analysis were accompanied by decreased variance around the mean, indicating that sites tend to be dominated by the functional type of the invading species.

Another surprising result was the prevalence of bimodal C4 abundance distributions across the entire dataset indicating that, at the plot scale, plant communities tend to be dominated largely by one functional type with few mixed C_3/C_4 intermediates. In fact, C_3 - and C_4 -dominated sites are largely separated in climate space, with overlap existing primarily where coexistence of functional types exists (Fig. S7). Whether or not these observations provide direct evidence for the existence of alternative stable states for different photosynthetic pathways is not known; however, the strongly suggestive nature of our data warrants future investigation of these patterns. For example, temporal demographic data from grasslands would be valuable for testing for alternative attractors. If grassland vegetation plots do exist in largely monocultures of either C_3 or C_4 , one likely explanation is that ecological conditions create feedbacks favoring a particular photosynthetic pathway creating non-linearity in species distributions. For example,

the direct and indirect promotion of C_4 grass by fire leads to nonlinearity in the climate responses the Eastern Temperate Forests and the highly bimodal abundance distribution observed (Fig. 4).

Soil type was a notably poor predictor of C₄ proportion, despite the frequent dominance of C₄ grasses on highly disturbed and low nutrient sites (Wilson & Tilman, 1993; Smith & Knapp, 1999); although, soil organic carbon was positively correlated with C₄ grasses in Eastern Temperate Forests. Climatic factors such as min/max temperature and precipitation, on the other hand, had stronger influence on grass distributions. In both Temperate Sierras and Southern Semiarid Highlands, higher MAP was associated with decreased C₄ proportions, in agreement with work on the Hawaiian Islands (Pau et al. 2013, Still et al. 2013), and potentially indicating that either increased rain favored C₃ grasses or that low rainfall favored C₄ (Paruelo & Lauenroth, 1996). Maximum temperatures positively affected C₄ grasses in Temperate Sierras and North American Deserts, which suggests that high temperatures have a particular negative effect on C₃ grasses in these areas (e.g., von Fischer *et al.* 2008). Similarly, higher minimum temperatures in Eastern Temperate Forests and Temperate Sierras were accompanied by higher C₄ proportions, indicating that cold may indeed limit C₄ production (Long, 1999).

Conclusion

We used vegetation abundance data to determine that counting the number of months favoring C_4 grasses, based on a max crossover temperature of 27 °C, was the best predictor of C_4 abundance. Furthermore, seven of eight biogeographical regions of North America examined had distinct and non-linear relationships with proportion of C_4 grasses predicted from the number of months favoring C_4 indicating that separate ecological processes might contribute differently to

distributional patterns among regions. In particular, invasive species, tree cover, and fire had important and regionally distinct modifying effects on C₄ grass abundance. In our study of C₃ and C₄ grass distributions in North America we found that climate disequilibrium was commonplace and biogeographically distinct at large scales.

Acknowledgements The authors would like to thank Matthias Schmid, Florian Wickler, Simone Wahl for providing both R code and detailed discussion regarding the application of Boosted Beta Regression. Susan Carr kindly provided data for the southeastern United States. Additional thanks to Tom Morrison, Frances Morris, and Kathleen Quigley for discussions. DG was supported by National Science Foundation Graduate Research Fellowship under Grant No.

0907738 and National Evolutionary Synthesis Center (NESCent), NSF #EF-0905606.

508 509	References:
510	Araújo, M.B., Ferri-Yáñez, F., Bozinovic, F., Marquet, P.A., Valladares, F. & Chown, S.L.
511	(2013) Heat freezes niche evolution. Ecology Letters, 16, 1206–1219.
512	Araújo, M.B. & Pearson, R.G. (2005) Equilibrium of species' distributions with climate.
513	Ecography, 28 , 693–695.
514	Araújo, M.B. & Rozenfeld, A. (2014) The geographic scaling of biotic interactions. <i>Ecography</i> ,
515	In press.
516	Asner, G.P., Elmore, A.J., Olander, L.P., Martin, R.E. & Harris, A.T. (2004) Grazing systems,
517	ecosystem responses, and global change. Annual Review of Environment and Resources,
518	29 , 261–299.
519	Barnes, P.W., Tieszen, L.L. & Ode, D.J. (1983) Distribution, production, and diversity of C ₃ -
520	and C ₄ -dominated communities in a mixed prairie. Canadian Journal of Botany, 61, 741
521	751.
522	Brown, J.H. & Maurer, B.A. (1989) Macroecology: the division of food and space among
523	species on continents. Science, 243, 1145–1150.
524	Callaway, R.M. (1995) Positive interactions among plants. <i>The Botanical Review</i> , 61 , 306–349.
525	Chesson, P. (2000) Mechanisms of maintenance of species diversity. Annual Review of Ecology
526	and Systematics, 31 , 343–366.
527	Collatz, G.J., Berry, J.A. & Clark, J.S. (1998) Effects of climate and atmospheric CO ₂ partial
528	pressure on the global distribution of C ₄ grasses: present, past, and future. <i>Oecologia</i> ,
529	114 , 441–454.

530	D'Antonio, C.M. & Vitousek, P.M. (1992) Biological Invasions by Exotic Grasses, the
531	Grass/Fire Cycle, and Global Change. Annual Review of Ecology and Systematics, 23,
532	63–87.
533	Dillon, M.E., Wang, G., Garrity, P.A. & Huey, R.B. (2009) Thermal preference in <i>Drosophila</i> .
534	Journal of Thermal Biology, 34 , 109–119.
535	Edwards, E.J., Osborne, C.P., Stromberg, C.A.E., Smith, S.A. & C ₄ Grasses Consortium (2010)
536	The Origins of C ₄ Grasslands: Integrating Evolutionary and Ecosystem Science. Science,
537	328 , 587–591.
538	Edwards, E.J. & Still, C.J. (2008) Climate, phylogeny and the ecological distribution of C ₄
539	grasses. Ecology Letters, 11, 266–276.
540	Ehleringer, J.R. (1978) Implications of quantum yield differences on the distributions of C ₃ and
541	C ₄ grasses. <i>Oecologia</i> , 31 , 255–267.
542	Ehleringer, J.R., Cerling, T.E. & Helliker, B.R. (1997) C ₄ photosynthesis, atmospheric CO ₂ , and
543	climate. <i>Oecologia</i> , 112 , 285–299.
544	Epstein, H.E., Lauenroth, W.K., Burke, I.C. & Coffin, D.P. (1997) Productivity patterns of C ₃
545	and C ₄ functional types in the US Great Plains. <i>Ecology</i> , 78 , 722–731.
546	Fargione, J. & Tilman, D. (2005) Niche differences in phenology and rooting depth promote
547	coexistence with a dominant C ₄ bunchgrass. <i>Oecologia</i> , 143 , 598–606.
548	Von Fischer, J.C., Tieszen, L.L. & Schimel, D.S. (2008) Climate controls on C ₃ vs. C ₄
549	productivity in North American grasslands from carbon isotope composition of soil
550	organic matter. Global Change Biology, 14, 1141-1155.
551	Guisan, A. & Rahbek, C. (2011) SESAM - a new framework integrating macroecological and
552	species distribution models for predicting spatio-temporal patterns of species

553	assemblages: Predicting spatio-temporal patterns of species assemblages. Journal of
554	Biogeography, 38 , 1433–1444.
555	Heckathorn, S.A., McNaughton, S.J. & Coleman, J.S. (1999) C ₄ Plants and Herbivory. C4 Plant
556	Biology, pp. 285–312. Academic Press, San Diego.
557	Horton, J.L. & Neufeld, H.S. (1998) Photosynthetic responses of <i>Microstegium vimineum</i> (Trin.)
558	A. Camus, a shade-tolerant, C ₄ grass, to variable light environments. <i>Oecologia</i> , 114 , 11–
559	19.
560	Long, S.P. (1999) Environmental Responses. C4 plant biology, pp. 313–373. Academic Press,
561	San Diego.
562	Maestre, F.T., Callaway, R.M., Valladares, F. & Lortie, C.J. (2009) Refining the stress-gradient
563	hypothesis for competition and facilitation in plant communities. Journal of Ecology, 97,
564	199–205.
565	Merow, C., Smith, M.J. & Silander, J.A. (2013) A practical guide to MaxEnt for modeling
566	species' distributions: what it does, and why inputs and settings matter. Ecography, no-
567	no.
568	Monson, R.K., Littlejohn Jr, R.O. & Williams III, G.J. (1983) Photosynthetic adaptation to
569	temperature in four species from the Colorado shortgrass steppe: a physiological model
570	for coexistence. <i>Oecologia</i> , 58 , 43–51.
571	Murphy, B.P. & Bowman, D.M.J.S. (2007) Seasonal water availability predicts the relative
572	abundance of C ₃ and C ₄ grasses in Australia. Global Ecology and Biogeography, 16,
573	160–169.
574	Ode, D.J., Tieszen, L.L. & Lerman, J.C. (1980) The Seasonal Contribution of C ₃ and C ₄ Plant
575	Species to Primary Production in a Mixed Prairie. <i>Ecology</i> , 61 , 1304–1311.

576	Osborne, C.P., Salomaa, A., Kluyver, T.A., Visser, V., Kellogg, E.A., Morrone, O., Vorontsova,
577	M.S., Clayton, W.D. & Simpson, D.A. (2014) A global database of C 4 photosynthesis in
578	grasses. New Phytologist, n/a–n/a.
579	Paruelo, J.M. & Lauenroth, W.K. (1996) Relative Abundance of Plant Functional Types in
580	Grasslands and Shrublands of North America. Ecological Applications, 6, 1212–1224.
581	Pau, S., Edwards, E.J. & Still, C.J. (2013) Improving our understanding of environmental
582	controls on the distribution of C ₃ and C ₄ grasses. Global Change Biology, 19, 184–196.
583	Peterson, D.W., Reich, P.B. & Wrage, K.J. (2007) Plant functional group responses to fire
584	frequency and tree canopy cover gradients in oak savannas and woodlands. Journal of
585	Vegetation Science, 18, 3–12.
586	Randerson, J.T., van der Werf, G.R., Collatz, G.J., Giglio, L., Still, C.J., Kasibhatla, P., Miller,
587	J.B., White, J.W.C., DeFries, R.S. & Kasischke, E.S. (2005) Fire emissions from C ₃ and
588	C_4 vegetation and their influence on interannual variability of atmospheric CO_2 and δ^{13}
589	CO ₂ . Global Biogeochemical Cycles, 19 , GB2019.
590	Ratnam, J., Bond, W.J., Fensham, R.J., Hoffmann, W.A., Archibald, S., Lehmann, C.E.R.,
591	Anderson, M.T., Higgins, S.I. & Sankaran, M. (2011) When is a "forest" a savanna, and
592	why does it matter?: When is a "forest" a savanna. Global Ecology and Biogeography,
593	20 , 653–660.
594	Sage, R.F. (2004) The evolution of C ₄ photosynthesis. <i>New Phytologist</i> , 161 , 341–370.
595	Sage, R.F., Wedin, D.A. & Li, M. (1999) The Biogeography of C ₄ Photosynthesis: Patterns and
596	Controlling Factors. C4 plant biology, pp. 313–373. Academic Press, San Diego.

597	Scheiter, S., Higgins, S.I., Osborne, C.P., Bradshaw, C., Lunt, D., Ripley, B.S., Taylor, L.L. &
598	Beerling, D.J. (2012) Fire and fire-adapted vegetation promoted C ₄ expansion in the late
599	Miocene. New Phytologist, 195, 653-666.
500	Scheiter, S., Langan, L. & Higgins, S.I. (2013) Next-generation dynamic global vegetation
501	models: learning from community ecology. New Phytologist, 198, 957–969.
502	Smith, M.D. & Knapp, A.K. (1999) Exotic plant species in a C ₄ -dominated grassland:
503	invasibility, disturbance, and community structure. Oecologia, 120, 605-612.
504	Still, C.J., Berry, J.A., Collatz, G.J. & DeFries, R.S. (2003a) Global distribution of C_3 and C_4
505	vegetation: Carbon cycle implications. Global Biogeochemical Cycles, 17, 1006.
606	Still, C.J., Berry, J.A., Ribas-Carbo, M. & Helliker, B.R. (2003b) The contribution of C_3 and C_4
507	plants to the carbon cycle of a tallgrass prairie: an isotopic approach. Oecologia, 136,
508	347–359.
509	Still, C.J., Pau, S. & Edwards, E.J. (2013) Land surface skin temperature captures thermal
510	environments of C_3 and C_4 grasses: Thermal niches and skin temperatures of C_3 and C_4
511	grasses. Global Ecology and Biogeography, 23, 286–296.
512	Taylor, S.H., Hulme, S.P., Rees, M., Ripley, B.S., Ian Woodward, F. & Osborne, C.P. (2010)
513	Ecophysiological traits in C ₃ and C ₄ grasses: a phylogenetically controlled screening
514	experiment. New Phytologist, 185, 780–791.
515	Taylor, S.H., Ripley, B.S., Martin, T., De-Wet, LA., Woodward, F.I. & Osborne, C.P. (2014)
516	Physiological advantages of C ₄ grasses in the field: a comparative experiment
517	demonstrating the importance of drought. Global Change Biology, In press.
518	Teeri, J.A. & Stowe, L.G. (1976) Climatic patterns and the distribution of C ₄ grasses in North
519	America. Oecologia, 23, 1–12.

620	Tieszen, L.L., Reed, B.C., Bliss, N.B., Wylie, B.K. & DeJong, D.D. (1997) NDVI, C_3 and C_4
621	Productions, and Distributions in Great Plains Grassland Land Cover Classes. Ecological
622	Applications, 7, 59–78.
623	Tilman, D. & Pacala, S. (1993) The Maintenance of Species Richness in Plant Communities.
624	Species Diversity in Ecological Communities, pp. 13–25. University of Chicago Press.
625	Veldman, J.W., Mattingly, W.B. & Brudvig, L.A. (2013) Understory plant communities and the
626	functional distinction between savanna trees, forest trees, and pines. Ecology, 94, 424-
627	434.
628	Weiher, E. & Keddy, P.A. (1995) Assembly rules, null models, and trait dispersion: new
629	questions from old patterns. Oikos, 159–164.
630	Wilson, S.D. & Tilman, D. (1993) Plant Competition and Resource Availability in Response to
631	Disturbance and Fertilization. <i>Ecology</i> , 74 , pp. 599–611.
632	Winslow, J.C., Hunt, E.R. & Piper, S.C. (2003) The influence of seasonal water availability on
633	global C ₃ versus C ₄ grassland biomass and its implications for climate change research.
634	Ecological Modelling, 163 , 153–173.
635 636	DIOGNETICIA
637 638	BIOSKETCH
639	Daniel M. Griffith's conducts plant ecological research in savanna and grassland ecosystems
640	with specific focus on the interactions of abiotic and biotic factors. His fieldwork is concentrated
641	on African grazing ecosystems.

Tables:

Table 1. Results from boosted beta regression analyses. Effect sized for variables are only presented if they were selected in the final model. A significant influence of space, which doesn't have a single effect size, is indicated as an asterisk. The symbols $(+\sigma^2)$ and $(-\sigma^2)$ indicates a positive and negative influences on the variance of C_4 proportion, indicated by the model as the inverse of beta distribution precision parameter (see methods). Only regions with models with significant mean effects are reported; these are Eastern Temperate Forest (ETF), Great Plains (GP), North American Deserts (NAD), Temperate Sierras (TS), and Southern Semiarid Highlands (SSH).

ETF	GP	NAD	TS	SSH
0.16	0.13	0.29	0.18	0.31
			-0.4	-0.18
-0.22	$(+\sigma^2)$		-0.07	-0.005
$0.1 (-\sigma^2)$		$-0.12 (-\sigma^2)$	-0.09	(-σ ²)
0.06				
$(+\sigma^2)$				
$(-\sigma^2)$				
0.17	$(+\sigma^2)$		$0.06 (+\sigma^2)$	
		0.03	0.02	
	*	*	*	
	0.16 -0.22 $0.1 (-\sigma^2)$ 0.06 $(+\sigma^2)$	0.16 0.13 -0.22 $(+\sigma^2)$ $0.1 (-\sigma^2)$ 0.06 $(+\sigma^2)$ 0.17 $(+\sigma^2)$	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$

Figure legends:

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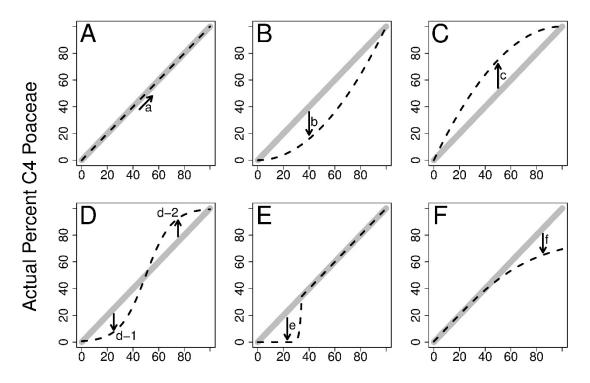
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Figure 1. Predicted patterns for the relationship between modeled and actual C₄ grass proportions. The gray line is the one-to-one "pure-physiology" prediction and the dotted black lines are alternative patterns. Capital letters identify the panel and lowercase letters identify the processes expected to produce the patterns observed. (A) A one-to-one relationship between expected and observed C_4 grasses would suggest that (a) physiology drives their distribution. (B) C₄ grasses might be represented below expected values in nature suggesting (b) C₃ competitive priority effects or disturbance favoring C₃. (C) C₄ grasses may be over represented and might imply (c) C₄ priority effects or disturbance favoring C₄. Further alternatives include (D) a sigmoidal relationship (d) where there are alternative states, thresholds, or opposing processes. (E) A lagged pattern where (e) the minimum threshold conditions for C₄ are not met. (F) An asymptotic curve that could suggest (f) storage effects buffering populations of C_3 species. Figure 2. The R² values from separate linear models for minimum, mean, and maximum temperature based crossover models across a range of potential values. The vertical line indicates the crossover temperature that maximizes the explanatory power of the model given the entire plot dataset. Figure 3. Map of the study extent, the conterminous USA, showing the biogeographic (Omernik) ecoregions analyzed (Datum: WGS84). Each region is associated with a plot using quantile regression to visualize the non-linear relationships (Appendix S2) between predicted and observed C₄ grasses. These graphs follow the framework developed in Fig 1, but use the

crossover temperature model as the x-axis. The gray line is the one-to-one, "climate equilibrium"

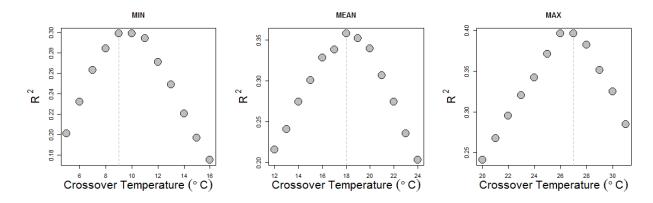
674 prediction. The data are represented with density curves within each bin on the x-axis (C4favored months). To visualize the median and upper limit of C₄ distributions, median and 95th 675 676 quantile regressions were fit in the R package 'quantreg' (Koenker 2013). 677 678 Figure 4. The Eastern Temperate Forest ecoregion was divided according to the historical range 679 of Longleaf pine. The figure shows the frequency distribution of C₄ grass proportion of the 680 Northeastern Temperate Forests (outside of the Longleaf pine range) as compared to the range of 681 Longleaf pine in the Southeastern USA. Each histogram has an inset depicting the results of a 682 path analysis testing the direct and indirect (mediated by tree cover) effects of fire on proportion 683 of C₄ grass. Significant regressions are labeled with standardized effect sizes. 684 685

Figures: 687 Figure 1.

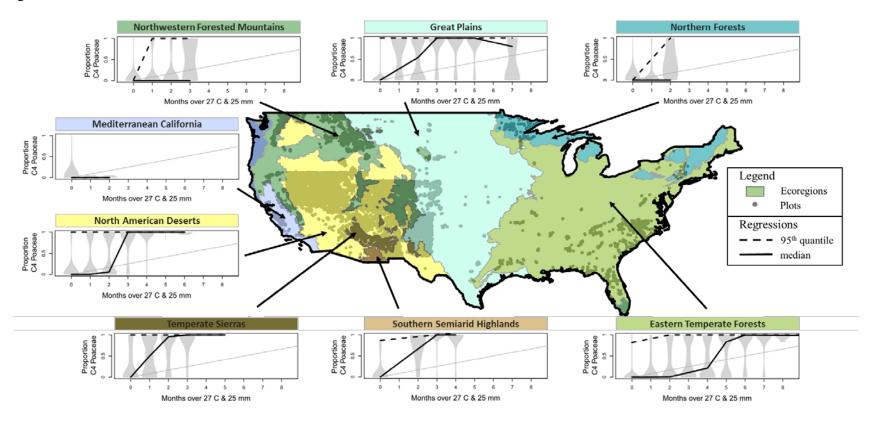


Predicted Percent C4 Poaceae

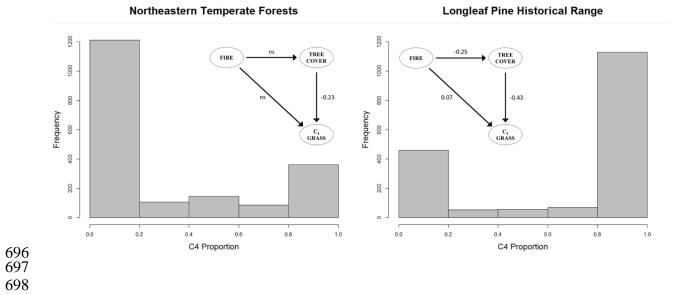
690 Figure 2.



692 Figure 3.



695 Figure 4.



699	SUPPORTING INFORMATION
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701	Appendix S1 Additional methods
702	Appendix S2 Additional results
703	Appendix S3 Details for path modeling
704	