# Supplementary Information: Impact of natural gas 

## extraction on PAH levels in ambient air

L. Blair Paulik ${ }^{\dagger}$, Carey E. Donald ${ }^{\dagger}$, Brian W. Smith ${ }^{\dagger}$, Lane G. Tidwell ${ }^{\dagger}$, Kevin A. Hobbie ${ }^{\dagger}$, Laurel Kincl $^{\ddagger}$, Erin N. Haynes ${ }^{\S}$, Kim A. Anderson ${ }^{\dagger} *$<br>${ }^{\dagger}$ Department of Environmental and Molecular Toxicology, Oregon State University, Corvallis, OR 97331<br>${ }^{\ddagger}$ College of Public Health and Human Sciences, Oregon State University, Corvallis, OR 97331<br>${ }^{\text {§ }}$ Department of Environmental Health, University of Cincinnati, Cincinnati, OH 45267

## Supplementary Information Table of Contents:

Chemical information - pg. 1
Passive sampler handling - pg. 1
Site description and sampling design - pg. 1
Volunteer training - pg. 1
Chemical analysis -pg. 1
Air concentration calculations - pg. 1
Sourcing ratios - pg. 2
PAHs used in comparison - pg. 2
Quantitative risk assessment calculations - pg. 2
Overview and discussion of exploratory principle components analysis (PCA) - pg. 3
Figure S1: Exploratory principle components analysis (PCA) - pg. 4
Table S1: List of QC and target analytes in GC/MS-MS method - pg. 5
Table S2: List of EPA Relative Potency Factors - pg. 8
Table S3: List of labels used in Figure S1 - pg. 9
References - pg. 11

## Chemical information

Single PAH standards were purchased from Sigma Aldrich, in St. Louis, MO, Chiron, in Trondheim, Norway, or Fluka (part of Sigma-Aldrich). PAH mixes were purchased from Accustandard, in New Haven, CT. Labeled compounds used as performance reference compounds (PRCs), laboratory surrogates, or instrument internal standards were obtained from either CDN Isotopes, in Pointe-Claire, Quebec, Canada, Cambridge Isotope Laboratories, in Tewksbury, MA, or Fisher Scientific in Pittsburgh, PA. All solvents were Optima-grade (from Fisher Scientific, Pittsburgh, PA) or equivalent, and all laboratory glassware and other tools were baked at $450^{\circ} \mathrm{C}$ for 12 hours and/or solvent-rinsed before use. Water used to clean LDPE was filtered through a D7389 purifier purchased from Barnstead International, in Dubuque, IA.

## Passive sampler handling

LDPE strips were cut from pre-sized polyethylene tubing that was approximately 2.7 cm wide. Each polyethylene strip was approximately 100 cm long and had a volume of $5.1 \mathrm{~cm}^{3}$. LDPE was dried under filtered vacuum in stainless steel kegs, from AEB Kegs in Delebio, Italy. TurboVap ${ }^{\circledR}$ evaporators were from Biotage, in Charlotte, NC.

## Site description and sampling design

A significant fraction of the Carroll County's residents earn their livings through farming. Carroll County also sits on both the Marcellus and Utica Shales. It has therefore been part of the natural gas boom occurring in the United States in recent years. The exact number of days that samplers were deployed ranged from 18 to 28. Welch Fluorocarbon, Inc. is in Dover, NH.

## Volunteer training

Volunteer landowners were trained in passive sampler handling, retrieval, and documentation by Oregon State University and University of Cincinnati community outreach specialists. Training included demonstration of retrieving samplers, practicing the technique, and time for discussion of questions and concerns. Volunteers were given handouts with step-by-step instructions of the sampler retrieval process. Handouts included a website with access to training videos, as well as contact information for OSU and UC trainers who would be available to answer questions.

## Chemical analysis

Agilent is located in Alpharetta, GA. The temperature profile in the GC/MS-MS analytical method was as follows: $60^{\circ} \mathrm{C}$ for 1 minute, increasing $40^{\circ} \mathrm{C}$ per minute to reach $180^{\circ} \mathrm{C}$, then increasing $3^{\circ} \mathrm{C}$ per minute to reach $230^{\circ} \mathrm{C}$, then increasing $1.5^{\circ} \mathrm{C}$ per minute to reach $235^{\circ} \mathrm{C}$, then increasing $15^{\circ} \mathrm{C}$ per minute to reach $280^{\circ} \mathrm{C}$, staying at $280^{\circ} \mathrm{C}$ for 10 minutes, then increasing $6^{\circ} \mathrm{C}$ per minute to reach $298^{\circ} \mathrm{C}$, and finally ramping up $16^{\circ} \mathrm{C}$ per minute to reach $350^{\circ} \mathrm{C}$ and stay there for 4 minutes. The dimensions of the Agilent Select PAH column were: 30 $\mathrm{m}, 0.25 \mathrm{~mm}, 0.15 \mu \mathrm{~m}$. Continuing calibration standards were run nominally every 10 samples, and/or at the end of the sample set. If a closing standard did not meet the criteria, samples were re-run after the standard was verified.

## Air concentration calculations

Vapor phase air concentrations were determined using an empirical uptake model. Sampling rates were derived by measuring PRC loss, as described in Huckins et al $2006{ }^{1}$. PRCs
allow for an accurate assessment of in situ uptake rates for a wide range of analytes in variable environmental conditions ${ }^{2-4}$. The uptake calculation does not make any assumptions about the analyte being at equilibrium, so this model was used for air concentration calculations for all PAHs. PRCs share similar physical and chemical properties with the target PAHs in this study and spanned a range of $\log K_{o a}$ values from 6.59 to $10.35^{3}$. Air concentrations $\left(C_{a}\right)$ of PAHs were determined using equation S1:

Eq. S1

$$
\mathrm{C}_{\mathrm{a}}=\frac{\mathrm{N}_{\text {analyte }}}{\mathrm{V}_{\mathrm{s}} \mathrm{~K}_{\mathrm{sa}}\left(1-\exp \left(-\frac{\mathrm{R}_{\mathrm{st}}}{\mathrm{~V}_{\mathrm{s}} \mathrm{~K}_{\mathrm{sa}}}\right)\right)}
$$

In equation $\mathrm{S} 1, \mathrm{C}_{\mathrm{a}}$ is the air concentration, $\mathrm{N}_{\text {analyte }}$ is the mass of the compound of interest present in the sampler, $\mathrm{V}_{s}$ is the sampler volume, $\mathrm{K}_{\mathrm{sa}}$ is the sampler-air partition coefficient, $\mathrm{R}_{\mathrm{s}}$ is the compound specific sampling rate, and $t$ is the duration of sampling. An analyte-specific $\mathrm{K}_{\mathrm{sa}}$ was calculated for each target PAH and PRC using a regression based on individual octanol-air partition coefficients ( $\mathrm{K}_{\text {oa }}$ ). Sampling rates $\left(\mathrm{R}_{\mathrm{s}}\right)$ of the PRCs were determined using equation S 2 :

Eq. S2

$$
R_{s}=-\frac{\ln \left(\frac{N}{N_{0}}\right)}{\mathrm{t}} \mathrm{~K}_{\mathrm{sa}} V_{\mathrm{s}}
$$

In equations S2 and S3, $N_{0}$ and $N$ are the mass of PRC present at the beginning and ending of the sampling period, respectively. The sampling rate $\left(R_{s}\right)$ for each analyte was calculated based on the $\mathrm{R}_{\mathrm{s}}$ of the PRC with the most similar $\mathrm{K}_{\mathrm{oa}}$. Eq S 3 uses compound class-specific modifiers (a) to compensate for compound-specific adjustments between the PRC and the target analyte.

$$
\text { Eq. S3 } \quad \mathrm{R}_{\mathrm{S}, \text { target analyte }}=\mathrm{R}_{\mathrm{S}, \mathrm{PRC}} * \frac{\alpha_{\text {analyte }}}{\alpha_{\mathrm{PRC}}}
$$

## Sourcing ratios

Petrogenic PAH sources are typically enriched in the more thermodynamically stable isomer ${ }^{5,6}$. In the two isomer pairs of PAHs used for PAH sourcing, phenanthrene and pyrene are the more thermodynamically stable isomers. Thus, a higher phenanthrene/anthracene, and a lower fluoranthene/pyrene ratio each indicate that the sample is predominantly petrogenic. A few samples had slightly pyrogenic signatures according to one ratio. However, more than one ratio should be used to confirm PAH source, as interpretation of values near the boundaries between sources can be less certain ${ }^{5}$.

## PAHs used in comparison

The 14 PAHs used in the comparison in Figure 2 were were Acenaphthylene, acenaphthene, fluorene, phenanthrene, anthracene, fluoranthene, pyrene, benzo[a]anthracene, chrysene, benzo[b]fluoranthene, benzo[k]fluoranthene, benzo[a]pyrene, benzo[g,h,i]perylene, and indeno[1,2,3-cd]pyrene. PAHs measured during the same season as the present study were used in this comparison where possible.

## Quantitative risk assessment calculations

There were 10 PAHs that were above the detection limits and had nonzero RPFs (see Table S2). Thus, these were the 10 PAHs that were used in the carcinogenic risk assessment. These 10 PAHs were benzo[a]pyrene, benzo[b]fluoranthene, cyclopenta[c,d]pyrene, benzo[j]fluoranthene, benzo[a]anthracene, chrysene, fluoranthene, indeno[1,2,3-c,d]pyrene, benzo[k]fluoranthene, and benzo[g,h,i]perylene.

Risk assessment was performed using equations from the EPA's 2009 Risk Assessment Guidance for Superfund ${ }^{7}$, using equations S4 and S5:

$$
\text { Eq. S4 } \quad \mathrm{EC}=\frac{(\mathrm{CA} \times \mathrm{ET} \times \mathrm{EF} \times \mathrm{ED})}{\mathrm{AT}}
$$

In equation $\mathrm{S} 4, \mathrm{EC}$ is the exposure concentration in $\mathrm{ng} / \mathrm{m}^{3}, \mathrm{CA}$ is the contaminant concentration in air in $\mathrm{ng} / \mathrm{m}^{3}$, ET is the exposure time in hours/day, EF is the exposure frequency in days/year, ED is the exposure duration in years, and AT is the averaging time. The AT includes the lifetime in years multiplied by 365 days/year and 24 hours/day.

$$
\text { Eq. S5 } \quad \text { ELCR }=I U R \times E C
$$

In equation S5, ELCR is excess lifetime cancer risk, IUR is inhalation unit risk, and EC is the exposure concentration from Equation S4. In this study, an IUR of $8.7 \times 10^{-5} \mathrm{ng} / \mathrm{m} 3$ was used. This is an IUR that was estimated for benzo[a]pyrene by the World Health Organization ${ }^{8}$. This was used because the U.S. EPA has no established an IUR for PAHs.

## Overview and discussion of exploratory principle components analysis (PCA)

Exploratory data analysis was performed using a Principle Component Analysis (PCA) biplot showing scores and loadings plots together. All PAH variables that were above detection limits were used, in addition to the distance to closest well variable. Only sites previously classified as close or far were included. The data were first square root transformed and then mean centered and scaled. PC1 and PC3 were chosen as graph axes because the resulting PCA graph displayed good delineation between close and far sites and suggested which variables contributed most to this 'clumping'. PCA was performed using Primer-E version 6.1.13. A list of labels used in the PCA plot and the PAHs they correspond to can be found in Table S3.

Figure S 1 shows that pyrene (p26) has the closest negative correlation with distance to the nearest active NGE well. This reinforces the results of the sourcing ratios in Figure 3. Given that pyrene is one of the PAHs used to indicate petrogenic signatures in the sourcing ratios, it is interesting that this is one of the main PAHs explaining the PAH data through PCA.

Figure S1 also shows samples in the close and far groups clumping separately. Looking closer at the clustering in the plot is also interesting. Samples in the far group are broken into two clusters of three samples. Far samples 1-3 cluster closer to the close samples, while far samples 4-6 cluster farther away from the rest of the samples. Interestingly, in the cluster that is closer to the close samples, two were located within 0.05 miles of heavily trafficked roads. This could potentially be a confounding factor.


Figure S1. Exploratory principle components analysis (PCA) using PAHs and distance to closest active NGE well treated as covariates. The graph only includes data for samples in the close $(\mathrm{n}=5)$ and far ( $\mathrm{n}=6$ ) distance groups.

Table S1: List of performance reference compounds (PRCs), internal standard (IS), surrogates, and target polycyclic aromatic hydrocarbons (PAH) in the GC/MS Triple Quad method used for PAH analysis in this study, with limits of detection (LOD) and limits of quantification (LOQ).

| PAH | Target, IS, PRC, or Surrogate? | CAS \# | $\begin{gathered} \text { LOD } \\ (\mathrm{ng} / \mathrm{mL}) \end{gathered}$ | $\begin{gathered} \text { LOQ } \\ (\mathrm{ng} / \mathrm{mL}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: |
| Fluorene-d10 | PRC | 81103-79-9 | . 33 | 1 |
| Pyrene-d10 | PRC | 1718-52-1 | 0.42 | 2.09 |
| Benzo[b]fluoranthene-d12 | PRC | 205-99-2 | 1.67 | 5 |
| Perylene-d12 | IS | 1520-96-3 | 1.67 |  |
| Naphthalene-d8 | Surrogate | 1146-65-2 | . 33 | 1 |
| Acenaphthylene-d8 | Surrogate | 93951-97-4 | . 33 | 1 |
| Phenanthrene-d10 | Surrogate | 1517-22-2 | 1.67 | 5 |
| Fluoranthene-d10 | Surrogate | 93951-69-0 | 1.67 | 5 |
| Chrysene-d12 | Surrogate | 1719-03-5 | 1.67 | 5 |
| Benzo[a]pyrene-d12 | Surrogate | 63466-71-7 | 1.67 | 5 |
| Benzo[ghi]perylene-d12 | Surrogate | 93951-66-7 | 1.67 | 5 |
| Naphthalene | Target | 91-20-3 | 1.04 | 5.20 |
| 2-Methylnaphthalene | Target | 91-57-6 | 0.70 | 3.50 |
| 1-Methylnaphthalene | Target | 90-12-0 | 0.28 | 1.39 |
| 2-Ethylnaphthalene | Target | 939-27-5 | 0.97 | 4.84 |
| 2,6-Dimethylnaphthalene | Target | 581-42-0 | 0.89 | 4.43 |
| 1,6-Dimethylnaphthalene | Target | 575-43-9 | 0.81 | 4.05 |
| 1,4-Dimethylnaphthalene | Target | 571-58-4 | 1.24 | 6.22 |
| 1,5-Dimethylnaphthalene | Target | 571-61-9 | 1.19 | 5.93 |
| 1,2-Dimethylnaphthalene | Target | 573-98-8 | . 94 | 4.70 |
| 1,8-Dimethylnaphthalene | Target | 569-41-5 | 0.83 | 4.15 |
| 2,6-Diethylnaphthalene | Target | 59919-41-4 | 0.81 | 4.06 |
| Acenaphthylene | Target | 208-96-8 | 2.33 | 11.65 |
| Acenaphthene | Target | 83-32-9 | 1.07 | 5.35 |
| Fluorene | Target | 86-73-7 | 0.79 | 3.97 |
| Dibenzothiophene | Target | 132-65-0 | 0.24 | 1.20 |
| Phenanthrene | Target | 85-01-8 | 0.46 | 2.31 |
| Anthracene | Target | 120-12-7 | 1.05 | 5.23 |
| 2-Methylphenanthrene | Target | 2531-84-2 | 0.39 | 1.93 |
| 2-Methylanthracene | Target | 613-12-7 | 0.47 | 2.36 |
| 1-Methylphenanthrene | Target | 832-69-9 | 1.06 | 5.32 |


| 9-Methylanthracene | Target | $779-02-2$ | 0.87 | 4.37 |
| :--- | :--- | :--- | :--- | :---: |
| 3,6-Dimethylphenanthrene | Target | $1576-67-6$ | 0.42 | 2.08 |
| 2,3-Dimethylanthracene | Target | $613-06-9$ | 0.34 | 1.71 |
| Fluoranthene | Target | $206-44-0$ | 0.54 | 2.72 |
| 9,10-Dimethylanthracene | Target | $781-43-1$ | 0.85 | 4.23 |
| Pyrene | Target | $129-00-0$ | 0.42 | 2.09 |
| Retene | Target | $483-65-8$ | 0.84 | 4.19 |
| Benzo[a]fluorene | Target | $238-84-6$ | 1.67 | 5 |
| Benzo[b]fluorene | Target | $243-17-4$ | 1.67 | 5 |
| Benzo[c]fluorene | Target | $205-12-9$ | 0.30 | 1.50 |
| 1-Methylpyrene | Target | $2381-21-7$ | 0.38 | 1.90 |
| Benz[a]anthracene | Target | $56-55-3$ | 0.75 | 3.77 |
| Cyclopenta[c,d]pyrene | Target | $27208-37-3$ | 0.53 | 2.67 |
| Triphenylene | Target | $217-59-4$ | 0.41 | 2.04 |
| Chrysene | Target | $218-01-9$ | 0.50 | 2.49 |
| 6-Methylchrysene | Target | $1705-85-7$ | 0.89 | 4.44 |
| 5-Methylchrysene | Target | $3697-24-3$ | 1.67 | 5 |
| Benzo[b]fluoranthene | Target | $205-99-2$ | 0.37 | 1.85 |
| 7,12-Dimethylbenz[a]anthracene | Target | $57-97-6$ | 0.94 | 4.71 |
| Benzo[k]fluoranthene | Target | $207-08-9$ | 0.53 | 2.63 |
| Benzo[j]fluoranthene | Target | $205-82-3$ | 0.56 | 2.79 |
| Benz[j]\&[e]aceanthrylene | Target | $202-33-5$ and | 1.67 | 5 |
| Benzo[e]pyrene | Target | $199-54-2$ |  | $07-2$ |
| Benzo[a]pyrene | Target | $50-32-8$ | 0.71 | 3.53 |
| Indeno(1,2,3-c,d)pyrene | Target | $193-39-5$ | 1.18 | 5.90 |
| Dibenzo[a,h]pyrene | Target | $53-70-3$ | 0.26 | 1.32 |
| Picene | Target | $213-46-7$ | 1.02 | 5.11 |
| Benzo[ghi]perylene | Target | $191-24-2$ | 0.74 | 3.72 |
| Anthanthrene | Target | $191-26-4$ | 0.34 | 1.71 |
| Naphtho[1,2-b]fluoranthene | $5385-22-8$ | 0.33 | 1.65 |  |
| Naphtho[2,3-j]fluoranthene | $205-83-4$ | 1.67 | 5 |  |
| Dibenzo[a,e]fluoranthene | $5385-75-1$ | 0.47 | 2.36 |  |
| Dibenzo[a,l]pyrene | $191-30-0$ | 0.48 | 2.41 |  |
| Naphtho[2,3-k]fluoranthene | Target | 1.67 | 5 |  |
| Naphtho[2,3-e]pyrene |  | 507 | 5 |  |


| Dibenzo[a,e]pyrene | Target | $192-65-4$ | 6.44 | 32.22 |
| :--- | :--- | :--- | :--- | :---: |
| Coronene | Target | $191-07-1$ | 0.70 | 3.49 |
| Dibenzo[e,l]pyrene | Target | $192-51-8$ | 1.67 | 5 |
| Naphtho[2,3-a]pyrene | Target | $196-42-9$ | 1.67 | 5 |
| Benzo[b]perylene | Target | $197-70-6$ | 1.67 | 5 |
| Dibenzo[a,i]pyrene | Target | $189-55-9$ | 1.42 | 7.10 |
| Dibenz[a,h]anthracene | Target | $189-64-0$ | 0.52 | 2.60 |

Table S2: "PAHs with final RPFs based on tumor bioassay data," from the U.S. EPA's 2010 Development of a Relative Potency Factor (RPF) Approach for Polycyclic Aromatic Hydrocarbon (PAH) Mixtures ${ }^{9}$.

| PAH | Relative Potency Factor |
| :--- | :---: |
| Anthanthrene | 0.4 |
| Anthracene | 0 |
| Benz[a]anthracene | 0.2 |
| Benz[b,c]aceanthrylene | 0.05 |
| Benzo[b]fluoranthene | 0.8 |
| Benzo[c]fluorene | 20 |
| Benz[e]aceanthrylene | 0.8 |
| Benzo[g,h,i]perylene | 0.009 |
| Benz[j]aceanthrylene | 60 |
| Benzo[j]fluoranthene | 0.3 |
| Benzo[k]fluoranthene | 0.03 |
| Benz[l]aceanthrylene | 5 |
| Chrysene | 0.1 |
| Cyclopenta[c,d]pyrene | 0.4 |
| Cyclopenta[d,e,f]chrysene | 0.3 |
| Dibenzo[a,e]fluoranthene | 0.9 |
| Dibenzo[a,e]pyrene | 0.4 |
| Dibenz[a,h]anthracene | 10 |
| Dibenzo[a,h]pyrene | 0.9 |
| Dibenzo[a,i]pyrene | 0.6 |
| Dibenzo[a,l]pyrene | 30 |
| Fluoranthene | 0.08 |
| Indeno[1,2,3-c,d]pyrene | 0.07 |
| Naphtho[2,3-e]pyrene | 0.3 |
| Phenanthrene | 0 |
| Pyrene | 0 |
| Benzo(a)pyrene | 1 |

Table S3: A list of labels used in the PCA plot in Figure S1, and the corresponding PAH names.

| PCA Label | PAH |
| :---: | :---: |
| p1 | Naphthalene |
| p2 | 2-Methylnaphthalene |
| p3 | 1-Methylnaphthalene |
| p4 | 2-Ethylnaphthalene |
| p5 | 2,6-Dimethylnaphthalene |
| p6 | 1,6-dimethylNaphthalene |
| p7 | 1,4-dimethylnaphthalene |
| p8 | 1,5-dimethylnaphthalene |
| p9 | 1,2-dimethylnaphthalene |
| p10 | 1,8-Dimethylnaphthalene |
| p11 | 2,6-Diethylnaphthalene |
| p12 | Acenaphthylene |
| p13 | Acenaphthene |
| p14 | Fluorene |
| p15 | Dibenzothiophene |
| p16 | Phenanthrene |
| p17 | Anthracene |
| p18 | 2-Methylphenanthrene |
| p19 | 2-Methylanthracene |
| p20 | 1-Methylphenanthrene |
| p21 | 9-Methylanthracene |
| p22 | 3,6-Dimethylphenanthrene |
| p23 | Fluoranthene |
| p24 | 2,3-Dimethylanthracene |
| p25 | 9,10-Dimethylanthracene |
| p26 | Pyrene |
| p27 | Retene |
| p28 | Benzo[a]fluorene |
| p29 | Benzo[b]fluorene |
| p30 | Benzo[c]fluorene |
| p31 | 1-Methylpyrene |
| p32 | Benzo[a]anthracene |
| p33 | Cyclopenta[cd]pyrene |
| p34 | Triphenylene |
| p35 | Chrysene |
| p36 | 6-Methyl chrysene |
| p37 | 5-Methylchrysene |
| p38 | Benzo [b] fluoranthene |
| p39 | 7,12-Dimethylbenz[a]anthracene |
| p40 | Benzo [k] fluoranthene |
| p41 | Benzo [j] fluoranthene |


| p42 | Benzo [e] pyrene |
| :--- | :--- |
| p43 | p44 |
| p45 | Benzo [a] pyrene |
| p46 | Indeno [1,2,3-c,d] pyrene |
| p47 | Dibenzo [a,h] anthracene |
| p48 | Benzo [a] chrysene |
| p49 | Benzo [g,h,i] perylene |
| p50 | Anthanthrene |
| p51 | Naphtho[1,2-b]fluoranthene |
| p52 | Naphtho[2,3-j]fluoranthene |
| p53 | Dibenzo [a,e] flouranthene |
| p54 | Dibenzo [a,l] pyrene |
| p55 | Naphtho[2,3-k]fluoranthrene |
| p56 | Naphtho[2,3-e]pyrene |
| p57 | Dibenzo [a,e] pyrene |
| p58 | Coronene |
| p59 | Dibenzo[e,l]pyrene |
| p60 | Naphtho[2,3-a]pyrene |
| p61 | Benzo [b] perylene |

## References

1. Huckins, J. N., Petty, Jimmie D., Booij, Kees. , Monitors of Organic Chemicals in the Environment Springer: New York, 2006.
2. Bartkow, M. E.; Jones, K. C.; Kennedy, K. E.; Holling, N.; Hawker, D. W.; Müller, J. F., Evaluation of performance reference compounds in polyethylene-based passive air samplers. Environ. Pollut. 2006, 144, (2), 365-370.
3. Huckins, J. N.; Petty, J. D.; Lebo, J. A.; Almeida, F. V.; Booij, K.; Alvarez, D. A.; Cranor, W. L.; Clark, R. C.; Mogensen, B. B., Development of the permeability/performance reference compound approach for in situ calibration of semipermeable membrane devices. Environmental Science \& Technology 2002, 36, (1), 85-91.
4. Söderström, H. S.; Bergqvist, P.-A., Passive air sampling using semipermeable membrane devices at different wind-speeds in situ calibrated by performance reference compounds. Environmental Science \& Technology 2004, 38, (18), 4828-4834.
5. Budzinski, H.; Jones, I.; Bellocq, J.; Pierard, C.; Garrigues, P., Evaluation of sediment contamination by polycyclic aromatic hydrocarbons in the Gironde estuary. Mar. Chem. 1997, 58, (1), 85-97.
6. Yunker, M. B.; Macdonald, R. W.; Vingarzan, R.; Mitchell, R. H.; Goyette, D.; Sylvestre, S., PAHs in the Fraser River basin: a critical appraisal of PAH ratios as indicators of PAH source and composition. Org. Geochem. 2002, 33, (4), 489-515.
7. EPA, U. S. Risk Assessment Guidance for Superfund Office of Superfund Remediation and Technology Innovation: Washington, D.C., 2009.
8. WHO Air quality guidelines for Europe; World Health Organization Regional Office for Europe: Copenhagen, Denmark, 2000.
9. EPA, U. S. Development of a Relative Potency Factor (RPF) Approach for Polycyclic Aromatic Hydrocarbon (PAH) Mixtures. ; Intregrated Risk Information Systems (IRIS):
Washington, D.C., 2010.
